Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.





## Effect of NO<sub>x</sub> on 1,3,5-trimethylbenzene (TMB) oxidation product distribution and particle formation

Julia Hammes<sup>1</sup>, Epameinondas Tsiligiannis<sup>1</sup>, Thomas F. Mentel<sup>1,2</sup>, Mattias Hallquist<sup>1</sup>\*

- <sup>1</sup>Department of Chemistry and Molecular Biology, University of Gothenburg, Gothenburg, Sweden
- <sup>2</sup>Institute of Energy and Climate Research, IEK-8: Troposphere, Forschungszentrum Jülich GmbH, Jülich, Germany
  - \* Correspondence to: <a href="mailto:hallq@chem.gu.se">hallq@chem.gu.se</a>

#### 10 Abstract

Secondary organic aerosol (SOA) represents a significant fraction of the tropospheric aerosol and its precursors are volatile organic compounds (VOC). Anthropogenic VOCs (AVOC) dominate the VOC budget in many urban areas with 1,3,5-trimethylbenzene (TMB) being among the most reactive aromatic AVOCs. TMB formed highly oxygenated organic molecules (HOM) in NO<sub>x</sub> free environment, which could contribute to new particle formation (NPF) depending on oxidation conditions were elevated OH oxidation enhanced particle formation. The experiments were performed in an oxidation flow reactor, the Go:PAM unit, under controlled OH oxidation conditions. By addition of NO<sub>x</sub> to the system we investigated the effect of NO<sub>x</sub> on particle formation and on the product distribution. We show that the formation of HOM and especially HOM accretion products, strongly varies with NO<sub>x</sub> conditions. We observe a suppression of HOM and particle formation with increasing NO<sub>x</sub>/ $\Delta$ TMB and an increase in the formation of organonitrates (ON) mostly at the expense of HOM accretion products. We propose reaction mechanisms/pathways that explain the formation and observed product distributions with respect to oxidation conditions. We hypothesize that, based on our findings from TMB oxidation studies, aromatic AVOCs may not contribute significantly to NPF under typical NO<sub>x</sub>/AVOC conditions found in urban atmospheres.

### 1 Introduction

Volatile organic compounds (VOC) are ubiquitous in the atmosphere and major precursors for secondary organic aerosol (SOA). SOA represents a dominant fraction of the tropospheric aerosol (Hallquist et al., 2009;Shrivastava et al., 2017;Gentner et al., 2017) and affects climate (Intergovernmental Panel on Climate, 2014) and health (WHO, 2016). Consequently research interest in SOA formation and properties is ranging from remote atmospheres (Ehn et al., 2012;Ehn et al., 2014;Kristensen et al., 2016) to densely populated and polluted environments (Chan and Yao, 2008;Hu et al., 2015;Guo et al., 2014;Hallquist et al., 2016). Following the study by Ehn et al. (Ehn et al., 2012), highly oxygenated organic molecules (HOM) with low volatilities, formed from the oxidation of biogenic volatile organic compounds (BVOCs), have attracted much research interest (Crounse et al., 2013;Ehn et al., 2014;Jokinen et al., 2014;Jokinen et al., 2015;Mentel et al., 2015;Yan et al., 2016;Berndt et al., 2016;Bianchi et al., 2019). These compounds have been shown to contribute to new particle formation (NPF) and to SOA growth (Ehn et al., 2014;Bianchi et al., 2016;Kirkby et al., 2016;Trostl et al., 2016;McFiggans et al., 2019),

Atmos. Chem. Phys. Discuss., https://doi.org/10.5194/acp-2019-395 Manuscript under review for journal Atmos, Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.



20



making them an important factor in the formation of atmospheric SOA. These oxidation products can be either described as HOM based on their high oxygen number (O> 6) (Bianchi et al., 2019) or as extremely low volatile organic compounds (ELVOC) based on their volatility (Donahue et al., 2012; Trostl et al., 2016). In this study, we will refer to oxidation products as HOM, because not all of the measured compounds may fulfill the criteria for ELVOC (Trostl et al., 2016; Kurtén et al., 2016). Gas phase autoxidation of alkylperoxy radicals (RO<sub>2</sub>) has been proposed as the formation mechanism for HOM (Crounse et al., 2013; Ehn et al., 2014; Jokinen et al., 2014). After the initial reaction of an oxidant with the VOC and subsequent addition of O<sub>2</sub> to the alkylradical (R), the produced RO<sub>2</sub> isomerizes via intra molecular H abstraction (H-shift). During this process a hydroperoxide group and a new R is formed. Additional O<sub>2</sub> addition and H-shift sequences can introduce large amounts of oxygen to the molecule and subsequently lower the vapour pressure. The chemistry of aromatic compounds is somewhat different compared to other VOCs as they can lose their aromaticity during the initial OH attack while they can retain the ring structure. Moreover, reaction products are more reactive than the parent compound. The produced RO<sub>2</sub> form an oxygen bridge, a bicyclic and potentially a tricyclic alkylradical (Molteni et al., 2018; Wang et al., 2017) before further oxidation processes open the ring structure. For 1,3,5-trimethylbenzene (TMB), emitted from combustion sources in the urban environment, Molteni et al. (2018) proposed a generalized reaction scheme for HOM formation after OH addition. According to their scheme, first generation alkylperoxy radicals with the general formula of C<sub>9</sub>H<sub>13</sub>O<sub>5-11</sub> were formed from the initial OH attack and subsequent H shift and O<sub>2</sub> addition sequences. A postulated second OH attack would result in propagating peroxy radical chains yielding radicals with the general formulas  $C_9H_{15}O_{7-11}$ .

Generally, the termination reaction of RO<sub>2</sub> (with a general m/z = x) with HO<sub>2</sub>, leads to the formation of hydro peroxides (m/z = x + 1) while termination reactions with other RO<sub>2</sub> can lead to the formation of a carbonyl (m/z = x - 17), a hydroxy group (m/z = x - 15) or dimers (m/z = 2x - 32) (Mentel et al., 2015; Jokinen et al., 2014; Rissanen et al., 2014). The propagation reaction of RO<sub>2</sub> with another RO<sub>2</sub> or NO also results in the formation of RO (m/z = x - 16). The RO can undergo internal H shift leading to the formation of an hydroxy group and subsequent a new peroxy radical which can continue the autoxidation sequences. The alkoxy step shifts of the observed m/z by 16 leads to overlap of different termination product sequences. During extensive oxidation the first generation products are subject to secondary chemistry increasing the numbers of products. Molteni et al. (Molteni et al., 2018) found several closed shell monomer products with the general formula  $C_9H_{12-16}O_{5-11}$  from the oxidation of TMB with OH. The formation of dimers with different number of H atoms in their study was explained by reactions of two first generation RO<sub>2</sub> radicals ( $C_{18}H_{26}O_{5-10}$ ), one first and one second generation RO<sub>2</sub> ( $C_{18}H_{28}O_{9-12}$ ) or two second generation RO<sub>2</sub> resulting in the dimer  $C_{18}H_{30}O_{11}$ . Although dimers have in general lower O:C ratios than monomers, they are expected to be less volatile due to higher molecular weight and more functional groups making them candidates to participate in nucleation processes (Kirkby et al., 2016).

Organonitrates (ON) are formed as soon as sufficient  $NO_x$  is present in the atmosphere. ON are highly important for the reactive nitrogen budget wherein the formation of highly functionalized organic nitrates can contribute significantly to secondary organic aerosol (Lee et al., 2016;Bianchi et al., 2017). In this study we refer to compounds that only consist of H, C and O as HOM monomers or HOM dimers and to N containing compounds as ONs.

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.



5

25



 $NO_x$  influences the oxidation of organics directly by changing oxidant levels (reducing or increasing OH, depending on the  $NO_x$  regime) and indirectly by influencing  $RO_2$  chemistry. In high  $NO_x$  environment such as urban areas the reaction of NO with  $RO_2$  radicals

$$RO_2 + NO \longrightarrow RONO2$$

$$\rightarrow RO + NO_2$$
(1)

can compete with the autoxidation mechanism (reaction 1 and 2). and thus potentially inhibit HOM while favouring ON formation (reaction 1). Although ON may have high O:C they differ from HOM as they contain at least one nitrogen atom. While HOM formation from BVOCs has been intensively studied, only few studies have been conducted focusing on the formation of HOM from anthropogenic aromatic volatile organic compounds (AVOCs) (Molteni et al., 2018;Wang et al., 2017). However, these studies indicate that AVOCs have a strong potential to form HOM under NO<sub>x</sub> free conditions and proposed that they may play a crucial role in NPF and particle growth of SOA in urban areas. Regarding SOA yields a number of smog chamber studies have been conducted in order to investigate the oxidation of TMB with OH radical under different NO<sub>x</sub> and aerosol seed conditions (Paulsen et al., 2005;Rodigast et al., 2017;Wyche et al., 2009;Sato et al., 2018;Huang et al., 2015). They reported that under higher NO<sub>x</sub>:VOC conditions, SOA yield is reduced compared to medium or lower NO<sub>x</sub>:VOC conditions. The NO<sub>x</sub>:VOC, in almost all studies, was < 1, apart from one experiment (Wyche et al., 2009), with NO<sub>x</sub>:VOC = 1.9 in which the SOA yield was 0.29, compared to yields up to 7.47 under lower NOx conditions.

In this study we investigate the oxidation of TMB in a laminar flow reactor, while different  $NO_x$  and OH conditions were applied. A nitrate chemical ionization atmospheric pressure interface time of flight mass spectrometer (CI-APi-TOF-MS) (Junninen et al., 2010; Jokinen et al., 2012) was used to monitor the oxidation product distribution. We show the formation of HOM and nitrate containing compounds with and without NOx added to the reaction system. Possible mechanisms leading to the formation of ON and suppression of particle formation are discussed.

#### 2 Materials and Methods

The measured HOM were generated using the laminar flow Gothenburg Potential Aerosol Mass reactor (Go:PAM), initially described by Watne et al. (Watne et al., 2018). The Go:PAM is a 100 cm long, 9.6 cm wide quartz glass cylinder which is irradiated over a length of 84 cm by two 30 W Phillips TUV lamps (254 nm); a schematic is shown in Figure 1. The OH radicals are produced inside Go:PAM by photolysing O<sub>3</sub> in the presence of water vapour. The O<sub>3</sub> is generated outside Go:PAM by photolyzing pure O<sub>2</sub> (UVP Pen-Ray® Mercury Lamps, 185 nm) and distributed in 3 L min<sup>-1</sup> particle free and humidified air (Milli – Q) over the reactor cross section. The VOC was introduced through a gravimetrically characterized diffusion source (see SI Figure 1) centrally at the top of the reactor with a flow of 8 L min–1. Flows were adjusted for a median residence time of 34 s in Go:PAM. A funnel shaped device is subsampling the centre part of the laminar flow to minimize wall effects on the sample. A condensation particle counter (CPC, 3775 TSI) was used to measure the number particle concentration in the sample flow. O<sub>3</sub> was monitored by a model 202 monitor (2B Technologies), relative humidity by a Vaisala HMP60 probe and NO<sub>x</sub> by a model 42i monitor (Thermo Scientific) over the course of the experiments. The OH exposure, over the residence time in the reactor, for NO<sub>x</sub> free conditions without added TMB was measured using SO<sub>2</sub> titration

Atmos. Chem. Phys. Discuss., https://doi.org/10.5194/acp-2019-395 Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.





(Teledyne T100) as described by Kang et al. (Kang et al., 2007). Oxidation products were measured with an Atmospheric Pressure interface High Resolution Time of Flight Mass Spectrometer (APi-TOF-MS, Aerodyne Research Inc. & Tofwerk AG) (Junninen et al., 2010; Jokinen et al., 2012) in connection with a A70 CI-inlet (Airmodus Ltd) (Eisele and Tanner, 1993). The CI inlet is a laminar flow inlet operated with a sheath flow of 20 L min<sup>-1</sup> containing NO<sub>3</sub><sup>-</sup> ions which are generated by ionizing HNO<sub>3</sub> using an <sup>241</sup>Am foil upstream in the inlet design. The sample stream from Go:PAM is introduced in the centre of the sheath flow at a rate of 8 L min<sup>-1</sup>. The NO<sub>3</sub><sup>-1</sup> ions are electrostatically pushed into the sample flow and form stable adducts with sample molecules as described by Ehn et al. (2012). The reaction time of oxidation products and NO<sub>3</sub> is a few hundred ms before being subsampled into the TOF-MS at 0.8 L min<sup>-1</sup> by a critical orifice. Differential pumping decreased the pressure from 103 mbar in the CI source to 10<sup>-6</sup> mbar in the TOF extraction region where HOM are detected as negatively charged clusters with NO<sub>3</sub>-.

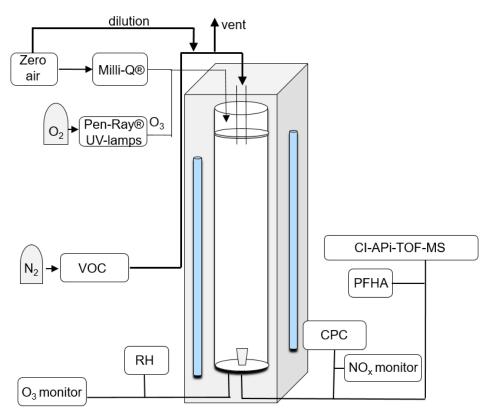


Figure 1: Schematics of the experimental setup with Go:PAM chamber connected to CI- APi-TOF-MS.

A kinetic box model was used to simulate the chemistry in the Go:PAM reactor. The core of the model were first described by Watne et al. (Watne et al., 2018). The model consists of 32 species and 68 reactions now including TMB chemistry partly from the MCM v3.3.1 (Jenkin et al., 2003) as well as proposed mechanisms and rate coefficients for NO<sub>2</sub> chemistry (Atkinson et al., 1992; Finlayson-Pitts, 1999) and highly oxygenated compounds (Ehn et al., 2014; Berndt et al., 2018; Zhao et al., 2018) (see SI 1). The photon flux used in the simulations was tuned to match measured decay of O<sub>3</sub> while an OH sink was added to match the observed OH exposure in the Atmos. Chem. Phys. Discuss., https://doi.org/10.5194/acp-2019-395 Manuscript under review for journal Atmos, Chem. Phys.

Discussion started: 17 May 2019







background experiment, i.e. without the addition of TMB. The model was run for all experiments with and without  $NO_x$ . Primarily, the modelled output on OH exposure for each experiments was used to interpret the results and for calculating the consumed TMB. However, broadly the modelling output was also used to understand the effects of changing experimental conditions on monomer, dimer and organo nitrate (ON) production. The experiments without  $NO_x$  were named 1-4 denoting the increase in OH exposure and experiments with  $NO_x$  were denoted according to their  $NO_x/\Delta TMB$  and high (H) and low (L) OH exposure as seen in Table 1.

#### 3 Results and Discussions

Table 1 summarises eight experiments where TMB has been oxidised by various amounts of OH using the Go:PAM unit. Generally, a high OH production, induced either by increased light exposure (two lamps) or elevated ozone concentration, resulted in new particles (e.g. exp 3 and 4) while addition of NO<sub>x</sub> reduced or suppressed the particle formation. The results from the kinetic model show that the amount of reacted TMB ranges from 5-30 ppb, depending on OH exposure (SI Figure 1). An overview of the oxidation product distribution measured with the CI-APi-TOF-MS for different conditions is shown in Figure 3 and Figure 4. The compounds were detected as nitrate clusters at  $m/z = \text{mass}_{\text{compound}} + 62$ . The spectra in Figure 2 and 3 show significant ion signals from oxygenated hydrocarbons retaining the 9 carbons from the original TMB with either even H numbers (closed shell) or odd H numbers (open shell) with limited amount of products from fragmentation, i.e. ions with less than nine C. C<sub>9</sub> compounds with an O/C ratio of 6/9 or higher were classified as HOM monomers with the general formula  $C_9H_{12-16}O_{6-11}$  in the mass range 280 – 360 m/z. Oxygenated hydrocarbons found in the range 460 – 560 m/z containing 18 C were classified as dimers with chemical formulas  $C_{18}H_{24-30}O_{10-16}$ . The monomer with the highest intensity detected was  $C_9H_{14}O_7$  m/z 296. The highest intensities among the dimers were  $C_{18}H_{26}O_{10}$ ,  $C_{18}H_{28}O_{11}$  and C<sub>18</sub>H<sub>28</sub>O<sub>12</sub> at m/z 464, m/z 482 and m/z 498, respectively. In addition to HOM monomers and dimers, nitrogen containing compounds were found as C<sub>9</sub> compounds with one or two N or C<sub>18</sub> compounds with one N. The nitrogen containing compounds were of the general formulas  $C_9H_{12-18}NO_{6-13}$ ,  $C_9H_{12-18}N_2O_{8-15}$  and  $C_{18}H_{18-24}NO_{6-10}$ . The dominating ON were  $C_9H_{13}NO_8$  at m/z 325,  $C_9H_{15}NO_{10}$  at m/z 359 and  $C_9H_{14}N_2O_{10}$  at m/z 372 respectively. In the experiments where NO<sub>x</sub> was added, the formation of ON compounds was increasing with the NO<sub>x</sub> concentration. In parallel, the levels of HOM monomers and dimers were reduced with NO<sub>x</sub> concentration, where dimers were stronger affected than monomers. Even if the fragmentation products were limited some fragmentation leading to less than 9 carbons could be observed. The most prominent fragments were assigned molecular formulas  $C_4H_7NO_7$  at m/z 243,  $C_4H_6O_{12}$  at m/z 246,  $C_5H_6O_{12}$  m/z 285 and  $C_6H_9NO_7$  at m/z 269. Some compounds with C numbers of 15 and 17 were detected in the range  $270 - 560 \, m/z$  but their contribution to the total signal was negligible (Figure 2).

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.





Table 1: Experimental conditions for experiments with 30 ppb TMB. Ozone and initial NOx concentration at time 0 are given in ppb and explicitly modelled OH exposure in molecules s cm<sup>-3</sup>. TMB reacted (Δ TMB) in ppb after a reaction time of 34 s and particle number concentration given in # cm<sup>-3</sup> after reaching steady state in Go:PAM. RH in all experiments was 38%.

#	[O <sub>3</sub> ] <sub>0</sub>	[NO <sub>x</sub> ] <sub>0</sub>	OH exposure	Δ TMB	NO <sub>x</sub> /ΔTMB	Particle number	Contribution of top 10 species (%)
1	~19	5	$3.5 \times 10^9$	5.4	0.9	-	28.6
2	~19	5	$7.1 \times 10^9$	9.9	0.5	-	29.8
3	~100	3	$3.8 \times 10^{10}$	26	0.1	$60 \pm 14$	38.6
4	~100	3	$2.1 \times 10^{11}$	30	0.1	$1610 \pm 217$	35.9
NOx9	~9	82	6.3×10 <sup>9</sup>	9	9.1	-	52.4
NOx3 <sub>L</sub>	~12	38	$7.9 \times 10^9$	11	3.5	-	42.3
NOx3 <sub>H</sub>	~100	79	$3.1 \times 10^{10}$	25	3.2	-	34.2
NOx1	~100	35	$9.1 \times 10^{10}$	30	1.2	$170 \pm 50$	30.5

The relative contribution of the different compound classes to the total assigned signal are shown in Figure 2. It is apparent that HOM monomers and dimers dominate in the experiments with low NO<sub>x</sub>. The contribution of the monomers to the total oxidation product signal ranges from 20.7-42.1% and dimers make 6.8 – 43.3% of the total, depending on experimental conditions. Dimer contributions are highest at high OH exposure (exp 3 and 4 with estimated OH exposure of  $3.8 \times 10^{10}$  and  $2.1 \times 10^{11}$  molecules s cm<sup>-3</sup>, respectively) and decrease with increasing NO<sub>x</sub>. Nitrated compounds dominated the spectra with contributions up to ~75% in the experiments with highest amount of NO<sub>x</sub>. Surprisingly, some nitrated compounds were also found in the experiments without added NO<sub>x</sub> which may stem from background NO contamination (~3-5ppb).

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.





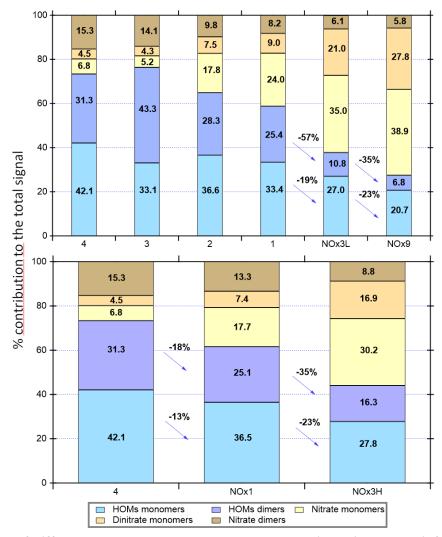


Figure 2: Overview of different compound groups to the total explained signal. The influence of increased NO<sub>x</sub>/VOC on the product distribution. Top panel illustrates the influence of a decrease of OH exposure (exp 4 - exp 1) and further decrease after adding NO<sub>x</sub> in exp NOx3L – exp NOx9. Dimers show a larger relative reduction than monomers with increasing NO<sub>x</sub>/VOC. Bottom panel shows the influence of increased NO<sub>x</sub>/VOC on the product distribution. Experiments 4, 3 and NOx1 resulted in particle formation.

Recently, Molteni et al. (2018) assigned 17 compounds making up 80% of the total detected signal for HOM oxidation products from the reaction of TMB with OH. Their compound with the highest fraction of the signal (24.2%) was the dimer  $C_{18}H_{26}O_8$ . This compound was not detected in our study. We did neither detected deprotonated compounds in the mass range  $270 - 560 \, \text{m/z}$  nor HOM monomers with 17 H nor compounds with O/C<0.55 which were found by Molteni et al. (2018). However, we do find 10 of the previously reported 17 monomers and dimers in our spectra. The oxidation product distributions in our experiments are in general term

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.





more diverse, i.e. we found more compounds with smaller yields compared to Molteni et al. (2018). In our experiments the highest 20 compounds together explain 46 - 63 % of the total signal with the individual highest oxidation products contributing only between 4.4 and 16%.

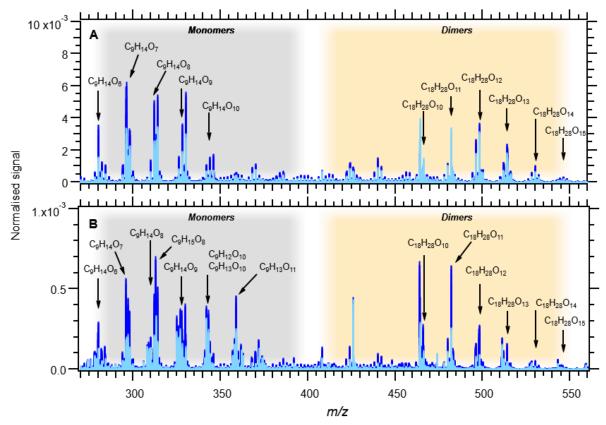


Figure 3: Mass spectra of all experiments without  $NO_x$ : 1, 2, 3, and 4. Panel A shows experiments 3 (light blue) and 4 (blue) with OH exposure of  $2.87 \times 10^{11}$  and  $3.47 \times 10^{11}$  molecules s cm<sup>-3</sup> respectively.Panel B shows experiment 1 (light blue) and 2 (blue) with of  $3.47 \times 10^{10}$  and  $7.62 \times 10^{10}$  molecules s cm<sup>-3</sup> respectively. The signal at m/z 426 is associated with the used mass calibrant PFHA. Note the ten times lower Normalised Signal scale in B.

The amount of TMB reacted after 34 s in the experiments ranges from 30 ppb (almost all) in experiment 4 to 5 ppb in experiments 1 and NOx9 (comp. Table 1 and Figure S1). The signal intensities of the HOM monomers are increasing with increasing OH exposure. Dimer compounds also increase with increasing OH and reach their highest levels in exp 3, resulting in the highest ratio of dimer to monomer (see Figure 2). Faster conversion of TMB will result in higher initial RO<sub>2</sub> levels enabling faster RO<sub>2</sub> + RO<sub>2</sub> (self-) reaction. The concentration profile of RO<sub>2</sub> in exp 1 and 2 is lower and more evenly spread out over the length of Go:PAM (Figure S1) and the influence of the RO<sub>2</sub> + RO<sub>2</sub> reaction will be less in these experiments compared to exp 3 and 4. Enhanced OH exposure does

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.



10



not only affect the monomer/dimer ratio but also the total amount of compounds measured. Observed compounds were 10 times higher in the high OH exposure exp 3, 4, compared to exp 1, 2. This was also valid for the experiments with added  $NO_x$  (NOx1 and  $NOx3_H$  vs  $NOx3_L$  and NOx9). Under high OH exposure (exp 4) the major HOM monomer is  $C_9H_{14}O_7$  with a contribution of 5.5 % to the total explained signal, followed by  $C_9H_{16}O_9$  and  $C_9H_{14}O_8$ , contributing 4.6 and 4.5 % respectively.

At the lowest OH exposure (exp 1) the major signals comprise the monomer  $C_9H_{12}O_{10}$  and the dimers  $C_{18}H_{26}O_{10}$  and  $C_{18}H_{28}O_{11}$ , contributing 4.4, 4.0 and 3.4 % to the total. These dimer compounds,  $C_{18}H_{26}O_{10}$  and  $C_{18}H_{28}O_{11}$ , are the highest signals in exp 2 and 3. Two open shell species were found among the larger signals in exp 1 and 2 (see Figure 2):  $C_9H_{15}O_8$  and  $C_9H_{15}O_7$ .

The number of H atoms is a characteristic for HOM monomers and dimers. An overview of different oxidation product generations is given in Table 2. We observe HOM monomers with 12-16 H and all compounds with an even number of H are closed shell products. Compounds with an uneven number of H are open shell molecules (radicals). Following the proposed termination scheme by Mentel et al. (2015) compounds with 12 H can be identified as first generation monomers (terminated from  $C_9H_{13}O_x$  radicals) and compounds with 16 H as second generation monomers (terminated from  $C_9H_{15}O_x$  radicals).  $C_9H_{14}O_x$  can originate from either  $C_9H_{13}O_x$  or  $C_9H_{15}O_x$ . The RO<sub>2</sub> radicals  $C_9H_{13,15}O_x$  can form dimers likely via the reaction

$$HOM - RO_2 + HOM - RO_2 \rightarrow HOM - dimer + O_2$$
 (3)

Such reaction is possible between any HOM-RO<sub>2</sub>, first or second generation (and other C<sub>9</sub> peroxy radicals with sufficient high abundance). Two first generation HOM-RO<sub>2</sub> will result in  $C_{18}H_{26}O_x$  dimers while one first and one second generation HOM-RO<sub>2</sub> produce a  $C_{18}H_{28}O_x$  and dimerization of two second generation HOM-RO<sub>2</sub> will result in dimers of the formula  $C_{18}H_{30}O_x$ .

A closer examination of the contribution of different HOM generations in Table 2 shows that first generation monomers with 12 H are showing the highest contribution in experiments with low OH exposure (exp 1, 2) while monomers with 14 H gain importance with higher OH exposure. The second generation monomers with 16 H dominate in experiments with the highest OH exposure (exp 3, 4). The dimer population with 28 H has a larger fraction of the total signal at higher OH exposures compared to dimers with 26 H. Dimer population with 30 H is generally lower than other dimers but has the highest fraction in exp 4. The overall dimer fraction of up to 43% in this study (Figure 2) is similar to the dimer fraction of 40% reported by Molteni et al. (2018). However the relative contributions of monomer and dimer generations differ. We find higher contributions of  $H_{12}$  monomers (up to 11%) and a higher contribution of  $H_{28}$  dimer (up to 12%) under our experimental conditions. The contribution of  $H_{14}$  monomers and  $H_{26}$  dimers is significantly less compared to Molteni et al. (2018).

Increasing OH exposure promotes second OH attacks on oxidation products leading to the observed reduction of first generation products  $(C_9H_{12}O_x)$  as well as increase of second generation products  $(C_9H_{14}O_x)$  and  $C_9H_{16}O_x$ . The increased oxidation degree can also explain the formation of dimers  $(C_18H_{28}O_x)$  from first and second generation  $RO_2$   $(C_9H_{13}O_x)$  and  $C_9H_{15}O_x$  at higher OH exposure and the increase in second generation dimers  $(C_9H_{15}O_x + C_9H_{15}O_x) + C_18H_{30}O_x)$ . Open shell species are observed as first generation  $RO_2$   $(C_9H_{13}O_x)$  and have a higher contribution lower OH exposures. Second generation  $RO_2$   $(C_9H_{15}O_x)$  have the highest contribution in exp 3. At the highest OH exposures in exp 4, the contribution of  $(C_9H_{15}O_x)$  is reduced at the expense of  $C_9H_{14}O_x$  and  $C_9H_{16}O_x$  - HOM and dimers.

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.





Table 2: Contribution of oxidation product families to the total signal between  $270 - 560 \, m/z$ 

Compound family	1	2	3	4	NOx9	NOx3 <sub>L</sub>	NOx3 <sub>H</sub>	NOx1
C <sub>9</sub> H <sub>12</sub> O <sub>x</sub>	11.3	7.8	3.5	5.4	8.5	9.3	5.9	5.4
$C_9H_{13}O_x$	5.7	4.5	3.3	2.1	6.0	7.0	5.6	4.2
C <sub>9</sub> H <sub>14</sub> O <sub>x</sub>	8.3	10.8	9.9	17.4	4.3	6.6	8.0	13.0
C <sub>9</sub> H <sub>15</sub> O <sub>x</sub>	5.6	7.2	9.4	4.1	2.4	3.3	4.6	4.7
$C_9H_{16}O_x$	4.8	8.0	7.7	14.5	1.5	2.5	4.8	10.3
$C_{18}H_{26}O_x$	8.5	9.1	8.5	9.3	0.8	2.0	2.7	6.0
$C_{18}H_{28}O_x$	7.1	9.9	9.3	11.0	0.4	1.2	2.6	6.9
$C_{18}H_{30}O_x$	0.7	1.0	2.3	2.6	0.4	0.4	0.9	1.7
C <sub>9</sub> H <sub>13</sub> NO <sub>x</sub>	6.2	4.7	0.6	0.6	26.8	17.1	10.1	4.8
C <sub>9</sub> H <sub>15</sub> NO <sub>x</sub>	6.4	5.6	0.7	0.9	3.7	10.3	14.5	8.5
$C_9H_{14}N_2O_x$	2.1	1.7	1.0	1.1	18.7	11.8	7.5	1.9

In this study the contribution of  $C_{18}H_{28}O_x$  shows that both first and second generation HOM-RO<sub>2</sub> were dimerising. The kinetics of dimer formation, if produced from RO<sub>2</sub> self-reaction, depends on the square of [HOM-RO<sub>2</sub>] and their relative importance will increase with the RO<sub>2</sub> concentration. Increased local RO<sub>2</sub> concentrations (in the first part of Go:PAM) would explain the increase of dimers with increasing OH exposure.

Substantial particle formation was observed in exp 4, under the highest OH exposure. Although the amount of reacted TMB in exp 3 and 4 is similar (26 and 30 ppb respectively), significant particle formation was not observed under the conditions of exp 3. The rate at which new particle formation (nucleation) occurs is related to the chemical composition and concentration of the nucleating species (McGraw and Zhang, 2008). After reaching the critical nucleus, particle growth becomes spontaneous in the presence of condensable vapour. Apparently, the local concentration of nucleating species or condensable vapour was not high enough in exp 3 to yield large numbers of particles, compared to exp 4. According to recent studies (Ehn et al., 2014;Trostl et al., 2016;Mohr et al., 2017;McFiggans et al., 2019) dimers play an important role in new particle formation. Mohr et al. (2017) found decreased levels of gas phase dimers in ambient air during NPF events, which is in line with our observations of lower dimer levels in the presence of particles in exp 4, compared to exp 3. A large enough concentration of low volatility dimers obviously helps forming critical nucleii that then grow by condensation. Note that the newly formed particles will provide an additional sink for dimers and thus reduce their presence as observable gas phase products at the end of the flow reactor.

5

Atmos. Chem. Phys. Discuss., https://doi.org/10.5194/acp-2019-395 Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.





#### Influence of NOx

In the experiments NOx1, NOx3<sub>H</sub>, NOx3<sub>L</sub> and NOx9, the NO<sub>x</sub> levels were increased (Table 1). The presence of NO<sub>x</sub> gave nitrogen containing C<sub>9</sub> compounds with one or two N atoms and C<sub>18</sub> compounds with only one N atom, in addition to HOM monomers and HOM dimers. N-containing compounds were the dominating species in experiments with NO<sub>x</sub> (see Figure 2), except for the experiment NOx1 with the smallest amount of NO<sub>x</sub> added and a high OH exposure (NOx1). The amount of nitrated compounds increased with the amount of added NO<sub>x</sub> at the expense of HOM monomers and HOM dimers as illustrated by the arrows in Figure 2. The effect of the added NO<sub>x</sub> is attenuated at higher OH exposure. E.g., the nitrated monomers are reduced from 35.0 and 38.9% at low OH exposure down to 17.7 and 30.2% at high OH exposure. Under elevated NO<sub>x</sub> conditions nitrated species were found among the 10 compounds with the highest contribution to the respective total signal, (see the top-ten lists in SI Table 1 and Figure 4).

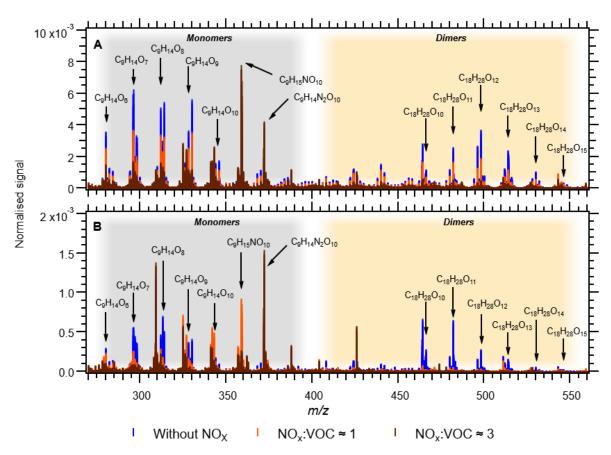


Figure 4: Comparison of mass spectra of HOM and nitrates. Panel A shows experiment 4 (blue), NOx1 (orange) and experiment NOx3<sub>H</sub> (brown). Panel B shows experiments 2 (blue), NOx3<sub>L</sub> (orange) and NOx9 (brown).

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.





For the two experiments with low OH exposure where the nitrated species dominated, seven out of the top ten species were nitrated with the highest signal arising from the di-nitrated compound ( $C_9H_{14}N_2O_{10}$ ). In these two experiments the top-ten compounds contributed the most to the observed signal (42.3 and 52.4%); most likely owing to the high fraction of nitrated species acting as radical chain termination products. Under highest NOx conditions, the high OH exposure experiment (NOx3<sub>H</sub>) has six nitrated compounds in the top-ten list with the dinitrated  $C_9H_{14}N_2O_{10}$  on the second rank. In experiment NOx1 only one nitrated species ( $C_9H_{15}NO_{10}$ ) is found in the top-ten list despite the elevated NOx conditions.

Dimer formation is drastically reduced in the presence of elevated  $NO_x$ . The  $NO_x$  effect on NPF was clear and high  $NO_x/\Delta TMB$  supresses NPF from TMB oxidation, a trend that was also observed by Wildt et al. (2014) and Lehtipalo et al. (2018) for NPF from monoterpene oxidation. Whenever the products were dominated by ON (NOx3<sub>L</sub>, NOx9 and NOx3<sub>H</sub>) particle formation was not observed. Reduced particle formation was observed in NOx1 compared to exp 1 which had a higher contribution of HOM. Owing to the importance of dimers for NPF, as reported by Lehtipalo et al. (2018), we suggest that the reaction of  $RO_2$  + NO resulting in ON is responsible for the observed reduced particle formation, because it competes with the dimer formation from  $RO_2$  +  $RO_2$ , This reaction is also reducing the contribution of HOM monomers and dimers with increasing  $NO_x/\Delta TMB$  and decreasing OH exposure from a total of 61% in exp NOx1 (OH exposure=9.1×10<sup>10</sup> molecules s cm<sup>-3</sup>) to 27.5% in NOx9 (OH exposure=6.3×10<sup>9</sup> molecules s cm<sup>-3</sup>). The reaction  $RO_2$  +  $RO_2$  or  $RO_2$ +HO<sub>2</sub>. The yield of ON from  $RO_2$  must be high or maybe close to unity based on the measurable increase of ON and the decrease of  $RO_2$  in the system (see Figure 7).

The contribution of HOM monomer generations follows the general trend observed in experiments without  $NO_x$ . The first generation HOM have higher contribution at low OH exposures in  $NOx3_L$  and NOx9 and the contribution of second generation HOM is higher the higher the OH exposure is. Similarly, the contribution of  $C_9H_{14}O_x$  is increasing with increasing OH exposure. Analogously to the families of HOM, different families of nitrates can be defined. Table 2 gives an overview of the nitrate families and how they contribute to the total signal in the experiments. At lowest OH exposures we find the highest contributions of first generation nitrates  $(C_9H_{13}NO_x)$  as well as di-nitrates  $(C_9H_{14}N_2O_x)$ . The contribution of second generation nitrates  $(C_9H_{15}NO_x)$  is increasing with increasing OH exposure and is highest in  $NOx3_H$ .

For the formation of observed first generation nitrates  $C_9H_{13}NO_{6-12}$  ( $C_9H_{13}NO_x$  in Table 2) we propose the reaction of a first generation (HOM-)RO<sub>2</sub> with NO following the pathway:

$$C_9H_{13}O_{5-11} + NO \rightarrow C_9H_{13}NO_{6-12}$$
 (4)

The precursor  $C_9H_{13}O_5$  is formed after two autoxidation steps and its termination reaction with NO results in  $C_9H_{13}NO_6$  which has only minor contribution.  $C_9H_{13}NO_{7\&8}$ , with higher contribution in exp NOx3<sub>L</sub> and NOx9 can be formed from the radicals  $C_9H_{13}O_6$  and  $C_9H_{13}O_7$  respectively. The even oxygen number in  $C_9H_{13}O_6$  indicates that the compound should have undergone a transformation to RO (via reaction with RO<sub>2</sub> or NO) and subsequent H-shift and further O<sub>2</sub> addition (Vereecken and Peeters, 2010;Mentel et al., 2015).

Second generation nitrates ( $C_9H_{15}NO_{8-12}$ ) can be formed after an additional OH attack (and introduction of an additional H) on a first generation (HOM) monomer, which explains the increase of these compounds with increasing OH exposure. The termination of the  $RO_2$  radical chain with NO (reaction 5) will then lead to the

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.



20

30



formation of the second generation nitrate. The formation of  $RO_2$  precursor species with lower O numbers, i.e.  $C_9H_{15}O_{7-8}$  likely stem from compounds terminated earlier in the radical chain process ( $C_9H_{14}O_{4-5}$ ), which do not fall in the typical HOM class (O/C  $\geq$  6/9). The reaction of a first generation nitrate with OH, followed by autoxidation could also possibly produce  $C_9H_{15}NO_{8-12}$  by terminating via hydroxyl (reaction 6) or hydroperoxy pathway (reaction 7).

$$C_9H_{15}O_{7-11} + NO \rightarrow C_9H_{15}NO_{8-12}$$
 (5)

$$C_9H_{14}NO_{8-12} + RO_2 \rightarrow C_9H_{15}NO_{8-12} + R=0$$
 (6)

$$C_9H_{14}NO_{8-12} + HO_2 \rightarrow C_9H_{15}NO_{8-12} + O_2$$
 (7)

For the formation of the most abundant di-nitrates of the formula  $C_9H_{14}N_2O_{8-12}$  (reaction 8), OH has to attack a nitrated compound  $C_9H_{13}NO_{6-10}$  and the RO<sub>2</sub> radical chain has to be terminated with NO:

$$C_9H_{14}NO_{7-11} + NO \rightarrow C_9H_{14}N_2O_{8-12}$$
 (8)

The precursor species  $C_9H_{14}NO_7$  would be in this case formed from the OH attack on the smallest possible nitrate  $C_9H_{13}NO_4$  (formed after one autoxidation step and NO termination).

It should be noted that the formation of peroxy nitrates via the reaction  $RO_2 + NO_2 \rightarrow RO_2NO_2$  with alkyl and acyl-RO<sub>2</sub> (PAN-like) cannot be ruled out as a potential formation mechanisms of nitrates.

The volatility of the produced ON seems to be too high (compared to dimers) to initiate NPF at the present concentrations as seen from the reduction in particle formation potential from exp 4 (HOM dominated, particle formation) to NOx1 (reduction in HOM and increase in ON, reduced particle formation) and  $NOx3_H$  (ON dominated, no particle formation)

Figure 5 shows the results for the evolution of HOM monomer, HOM Dimers and organic nitrates (ON) as calculated using the kinetic model described in detailed in SI. The model used has a very simplified scheme for TMB oxidation and subsequent  $RO_2$  chemistry. We implemented a few rate coefficients suggested in the literature in order to demonstrate how those compare with our experimental results. Rate coefficients for  $RO_2$  termination reactions from MCM in combination with rate coefficients of dimer formation by Berndt et al. (2018) (case 1 and 2 in SI 1) led to an overestimation of dimer compounds. Better representation of the observations were achieved by applying a) the rate coefficients proposed by Zhang et al. (2018) for dimer formation  $(2.0 \times 10^{-12} \text{ cm}^3 \text{ molecules}^{-1} \text{ s}^{-1})$  and  $1.0 \times 10^{-12} \text{ cm}^3$  molecules<sup>-1</sup> s<sup>-1</sup> for termination and RO formation with branching ratios of 0.4 and 0.6, respectively; b) the rate coefficient from Berndt et al. (2018) for ON formation from HOMRO<sub>2</sub> + NO (1.0×10<sup>-11</sup> cm<sup>3</sup> molecules<sup>-1</sup> s<sup>-1</sup>, branching ratio 0.3 for reaction 56).

In Figure 5, the calculated HOM dimer contribution are the sum of medium (produced from HOM-RO<sub>2</sub>+RO<sub>2</sub>), and highly oxidised dimers (produced from HOM-RO<sub>2</sub> + HOM-RO<sub>2</sub>) while the ON includes both organic nitrates (produced from NO+HOM-RO<sub>2</sub>) and peroxy nitrates (produced from HOM-RO<sub>2</sub>+NO<sub>2</sub>). The increased production of monomers calculated for exp 1 and 2 is in agreement with experimental results, where we observed a slightly larger contribution from monomers compared to dimers. Calculated concentrations for dimers are higher in exp 3 and 4, compared to monomers. Secondary chemistry (reaction of OH with the products and possible formation of a second generation of HOM or nitrates) was not taken into consideration in the model. For the experiment with  $NO_x$  (Figure 5, right panels) the modeled product distribution follows the general trend of monomer, dimer and nitrate that we observe in the exp  $NOx3_L$  and NOx9 with a general higher nitrate production

Atmos. Chem. Phys. Discuss., https://doi.org/10.5194/acp-2019-395 Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019 © Author(s) 2019. CC BY 4.0 License.



5



compared to monomers and dimers. For the  $NO_x$  experiments with high initial ozone, the model can reproduce the higher HOM monomer and dimer levels in NOx1 but slightly overestimates the contribution of dimers. Particle formation was observed in NOx1 which might explain the overestimation owing to missing condensation sink in the model. Modelling  $NOx3_H$  gives an overestimation of monomers and dimers.

In NOx3<sub>H</sub> modelled dimers start forming after ~15 sec. Almost all NO is converted to ON or NO<sub>2</sub> at this point and the reaction HOM-RO<sub>2</sub>+NO does not produce additional ON and the modelled levels of ON reach a plateau while contribution of HOM dimers can increase. For exp NOx1 and NOx3<sub>H</sub> the model is in better agreement with the measurements if only the HOM - dimer formation (from HOM-RO<sub>2</sub>+HOM-RO<sub>2</sub>) is taken into account, i.e. highly oxidized dimers excluding medium oxidized dimer (from HOM-RO<sub>2</sub> + RO<sub>2</sub>). Generally, the simplified model was able to support our analysis of the TMB chemistry as described above. Furthermore, it gave some support to the recently suggested mechanisms included in case 3 (e.g. Zhang et al. (2018))

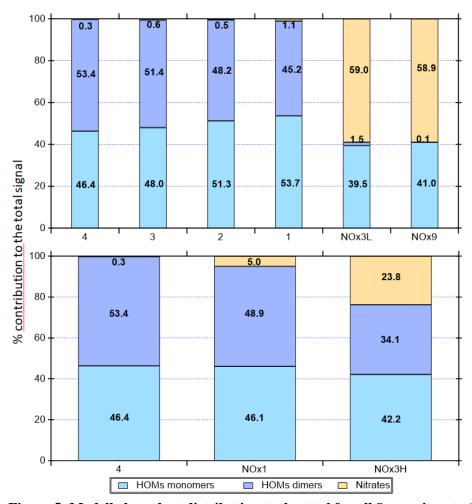


Figure 5: Modelled product distribution to the total for all 8 experiments (compare Fig. 2).

Atmos. Chem. Phys. Discuss., https://doi.org/10.5194/acp-2019-395 Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019 © Author(s) 2019. CC BY 4.0 License.





# **Atmospheric implication and Conclusion**

We have measured the formation of HOM monomers and dimers from OH initiated oxidation of TMB. The experiments with highest OH exposures lead to particle formation when NO<sub>x</sub> was not added. With increasing OH exposure and increased likelihood of a second OH attack, we observe a higher contribution from second generation oxidation products and dimers in general. The latter is attributed to the increased RO<sub>2</sub> concentrations from the increased/fast TMB consumption by OH. The observed products in this study match what would be expected as termination products from previously proposed reaction mechanisms for HOM formation.

The addition of  $NO_x$  to simulate urban condition leads to the formation of ON in addition to HOM and a reduction in particle formation potential. We observe that the formation of ON is increasing with increasing  $NO_x/\Delta TMB$ , mostly at the expense of dimers. The presence of ONs, formed at the expense of dimers, can explain the decreased tendency for particle formation. We therefore suggest that the reaction HOM-RO<sub>2</sub>+NO competes with HOM-RO<sub>2</sub> self-reaction yielding primarily a reduction in dimer formation, which is responsible for the reduction in particle formation.

According to studies by Molteni et al. (2018) and Wang et al. (2018) HOM formation from AVOCs was observed and consequently AVOC-HOM were suggested as potential contributors to observed NPF in urban atmospheres. In our study, under NO<sub>x</sub> free conditions, we found several of previous identified HOM even if we did not fully agree with the identity and relative importance of all HOM. However, the oxidation in polluted environments will happen under elevated NO<sub>x</sub> levels and, as has been shown here this can lead to formation of ON instead of HOM and subsequently a reduction in NPF potential. We conclude that for interpretation of NPF from aromatics in urban areas care should be taken and the OH exposure, NO<sub>x</sub> levels and RO<sub>2</sub> concentrations need to be considered in details since they will largely determine if the HOM-RO<sub>2</sub>+NO can compete with reactions yielding HOM, and especially HOM dimers.

Data availability: The data set is available upon request by contacting Mattias Hallquist (hallq@chem.gu.se).

Competing Interests: The authors declare that they have no conflict of interest.

Author contribution: Julia Hammes analysed the data and prepared the manuscript with contributions from all co-authors. The experiments were designed by Julia Hammes and Epameinondas Tsiligiannis. Epameinondas Tsiligiannis developed the model code and performed the simulations.

*Acknowledgements*: The research presented is a contribution to the Swedish strategic research area ModElling the Regional and Global Earth system, MERGE. This work was supported by the Swedish Research Council (grant numbers 2014-05332; 2013-06917) and Formas (grant number 2015-1537)

35

15

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.



20

25

40

50



#### 5. References:

- Atkinson, R., Baulch, D. L., Cox, R. A., Hampson, R. F., Kerr, J. A., and Troe, J.: Evaluated Kinetic and Photochemical Data for Atmospheric Chemistry Supplement-Iv Iupac Subcommittee on Gas Kinetic Data Evaluation for Atmospheric Chemistry, J Phys Chem Ref Data, 21, 1125-1568, Doi 10.1063/1.555918, 1992.
- Berndt, T., Richters, S., Jokinen, T., Hyttinen, N., Kurten, T., Otkjaer, R. V., Kjaergaard, H. G., Stratmann, F., Herrmann, H., Sipila, M., Kulmala, M., and Ehn, M.: Hydroxyl radical-induced formation of highly oxidized organic compounds, Nat Commun, 7, 13677, 10.1038/ncomms13677, 2016.
- Bianchi, F., Trostl, J., Junninen, H., Frege, C., Henne, S., Hoyle, C. R., Molteni, U., Herrmann, E., Adamov, A., Bukowiecki, N., Chen, X.,
  Duplissy, J., Gysel, M., Hutterli, M., Kangasluoma, J., Kontkanen, J., Kurten, A., Manninen, H. E., Munch, S., Perakyla, O., Petaja,
  T., Rondo, L., Williamson, C., Weingartner, E., Curtius, J., Worsnop, D. R., Kulmala, M., Dommen, J., and Baltensperger, U.: New
  particle formation in the free troposphere: A question of chemistry and timing, Science, 352, 1109-1112, 10.1126/science.aad5456,
  2016.
- Bianchi, F., Garmash, O., He, X., Yan, C., Iyer, S., Rosendahl, I., Xu, Z., Rissanen, M. P., Riva, M., Taipale, R., Sarnela, N., Petäjä, T.,
  Worsnop, D. R., Kulmala, M., Ehn, M., and Junninen, H.: The role of highly oxygenated molecules (HOMs) in determining the composition of ambient ions in the boreal forest, Atmos Chem Phys, 17, 13819-13831, 10.5194/acp-17-13819-2017, 2017.
  - Bianchi, F., Kurten, T., Riva, M., Mohr, C., Rissanen, M. P., Roldin, P., Berndt, T., Crounse, J. D., Wennberg, P. O., Mentel, T. F., Wildt, J., Junninen, H., Jokinen, T., Kulmala, M., Worsnop, D. R., Thornton, J. A., Donahue, N., Kjaergaard, H. G., and Ehn, M.: Highly Oxygenated Organic Molecules (HOM) from Gas-Phase Autoxidation Involving Peroxy Radicals: A Key Contributor to Atmospheric Aerosol, Chem Rev, 119, 3472-3509, 10.1021/acs.chemrev.8b00395, 2019.
  - Chan, C. K., and Yao, X.: Air pollution in mega cities in China, Atmos Environ, 42, 1-42, 10.1016/j.atmosenv.2007.09.003, 2008.
  - Crounse, J. D., Nielsen, L. B., Jorgensen, S., Kjaergaard, H. G., and Wennberg, P. O.: Autoxidation of Organic Compounds in the Atmosphere, Journal of Physical Chemistry Letters, 4, 3513-3520, 10.1021/jz4019207, 2013.
  - Donahue, N. M., Kroll, J. H., Pandis, S. N., and Robinson, A. L.: A two-dimensional volatility basis set Part 2: Diagnostics of organicaerosol evolution, Atmos Chem Phys, 12, 615-634, 10.5194/acp-12-615-2012, 2012.
  - Ehn, M., Kleist, E., Junninen, H., Petäjä, T., Lönn, G., Schobesberger, S., Dal Maso, M., Trimborn, A., Kulmala, M., Worsnop, D. R., Wahner, A., Wildt, J., and Mentel, T. F.: Gas phase formation of extremely oxidized pinene reaction products in chamber and ambient air, Atmos Chem Phys, 12, 5113-5127, 10.5194/acp-12-5113-2012, 2012.
- Ehn, M., Thornton, J. A., Kleist, E., Sipila, M., Junninen, H., Pullinen, I., Springer, M., Rubach, F., Tillmann, R., Lee, B., Lopez-Hilfiker, F., Andres, S., Acir, I.-H., Rissanen, M., Jokinen, T., Schobesberger, S., Kangasluoma, J., Kontkanen, J., Nieminen, T., Kurten, T., Nielsen, L. B., Jorgensen, S., Kjaergaard, H. G., Canagaratna, M., Maso, M. D., Berndt, T., Petaja, T., Wahner, A., Kerminen, V.-M., Kulmala, M., Worsnop, D. R., Wildt, J., and Mentel, T. F.: A large source of low-volatility secondary organic aerosol, Nature, 506, 476-479, 10.1038/nature13032, 2014.
- Eisele, F. L., and Tanner, D. J.: Measurement of the gas phase concentration of H2SO4and methane sulfonic acid and estimates of H2SO4production and loss in the atmosphere, Journal of Geophysical Research: Atmospheres, 98, 9001-9010, 10.1029/93jd00031, 1993
  - Finlayson-Pitts, B. a. P., J.: Chemistry of the Upper and Lower Atmosphere, 1999.
  - Gentner, D. R., Jathar, S. H., Gordon, T. D., Bahreini, R., Day, D. A., El Haddad, I., Hayes, P. L., Pieber, S. M., Platt, S. M., de Gouw, J., Goldstein, A. H., Harley, R. A., Jimenez, J. L., Prevot, A. S. H., and Robinson, A. L.: Review of Urban Secondary Organic Aerosol Formation from Gasoline and Diesel Motor Vehicle Emissions, Environmental science & technology, 51, 1074-1093, 10.1021/acs.est.6b04509, 2017.
  - Guo, S., Hu, M., Zamora, M. L., Peng, J., Shang, D., Zheng, J., Du, Z., Wu, Z., Shao, M., Zeng, L., Molina, M. J., and Zhang, R.: Elucidating severe urban haze formation in China, Proc Natl Acad Sci U S A, 111, 17373-17378, 10.1073/pnas.1419604111, 2014.
- Hallquist, M., Wenger, J. C., Baltensperger, U., Rudich, Y., Simpson, D., Claeys, M., Dommen, J., Donahue, N. M., George, C., Goldstein,
   A. H., Hamilton, J. F., Herrmann, H., Hoffmann, T., Iinuma, Y., Jang, M., Jenkin, M. E., Jimenez, J. L., Kiendler-Scharr, A.,
   Maenhaut, W., McFiggans, G., Mentel, T. F., Monod, A., Prevot, A. S. H., Seinfeld, J. H., Surratt, J. D., Szmigielski, R., and Wildt,
   J.: The formation, properties and impact of secondary organic aerosol: current and emerging issues, Atmos. Chem. Phys., 9, 5155-5236, 2009.
  - Hallquist, M., Munthe, J., Hu, M., Wang, T., Chan, C. K., Gao, J., Boman, J., Guo, S., Hallquist, A. M., Mellqvist, J., Moldanova, J., Pathak, R. K., Pettersson, J. B. C., Pleijel, H., Simpson, D., and Thynell, M.: Photochemical smog in China: scientific challenges and implications for air-quality policies, National Science Review, 3, 401-403, 10.1093/nsr/nww080, 2016.
    - Hu, M., Guo, S., Peng, J.-f., and Wu, Z.-j.: Insight into characteristics and sources of PM2.5in the Beijing–Tianjin–Hebei region, China, National Science Review, 2, 257-258, 10.1093/nsr/nwv003, 2015.
- Huang, M., Lin, Y., Huang, X., Liu, X., Guo, X., Hu, C., Zhao, W., Gu, X., Fang, L., and Zhang, W.: Experimental study of particulate products for aging of 1,3,5–trimethylbenzene secondary organic aerosol, Atmos Pollut Res, 6, 209-219, 10.5094/apr.2015.025, 2015.

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.



5

35

40



- Intergovernmental Panel on Climate, C.: Climate Change 2013 The Physical Science Basis, Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 1535 pp., 2014.
- Jokinen, T., Sipilä, M., Junninen, H., Ehn, M., Lönn, G., Hakala, J., Petäjä, T., Mauldin, R. L., Kulmala, M., and Worsnop, D. R.: Atmospheric sulphuric acid and neutral cluster measurements using CI-APi-TOF, Atmos Chem Phys, 12, 4117-4125, 10.5194/acp-12-4117-2012, 2012.
- Jokinen, T., Sipila, M., Richters, S., Kerminen, V. M., Paasonen, P., Stratmann, F., Worsnop, D., Kulmala, M., Ehn, M., Herrmann, H., and Berndt, T.: Rapid autoxidation forms highly oxidized RO2 radicals in the atmosphere, Angewandte Chemie (International ed. in English), 53, 14596-14600, 10.1002/anie.201408566, 2014.
- Jokinen, T., Berndt, T., Makkonen, R., Kerminen, V. M., Junninen, H., Paasonen, P., Stratmann, F., Herrmann, H., Guenther, A. B., Worsnop,
  D. R., Kulmala, M., Ehn, M., and Sipila, M.: Production of extremely low volatile organic compounds from biogenic emissions:
  Measured yields and atmospheric implications, Proc. Natl. Acad. Sci. U.S.A, 112, 7123-7128, 10.1073/pnas.1423977112, 2015.
  - Junninen, H., Ehn, M., Petäjä, T., Luosujärvi, L., Kotiaho, T., Kostiainen, R., Rohner, U., Gonin, M., Fuhrer, K., Kulmala, M., and Worsnop, D. R.: A high-resolution mass spectrometer to measure atmospheric ion composition, Atmos Meas Tech, 3, 1039-1053, 10.5194/amt-3-1039-2010, 2010.
- 15 Kang, E., Root, M. J., Toohey, D. W., and Brune, W. H.: Introducing the concept of Potential Aerosol Mass (PAM), Atmos Chem Phys, 7, 5727-5744, 10.5194/acp-7-5727-2007, 2007.
  - Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon, M., Yan, C., Almeida, J., Trostl, J., Nieminen, T., Ortega, I. K., Wagner, R., Adamov, A., Amorim, A., Bernhammer, A. K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A., Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J., Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kurten, A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Molteni,
- Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kurten, A., Laaksonen, A., Lehtipalo, K., Makhmutov, V., Mathot, S., Molteni, U., Onnela, A., Perakyla, O., Piel, F., Petaja, T., Praplan, A. P., Pringle, K., Rap, A., Richards, N. A., Riipinen, I., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sipila, M., Steiner, G., Stozhkov, Y., Stratmann, F., Tome, A., Virtanen, A., Vogel, A. L., Wagner, A. C., Wagner, P. E., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P., Zhang, X., Hansel, A., Dommen, J., Donahue, N. M., Worsnop, D. R., Baltensperger, U., Kulmala, M., Carslaw, K. S., and Curtius, J.: Ion-induced nucleation of pure biogenic particles, Nature, 533, 521-526, 10.1038/nature17953, 2016.
  - Kristensen, K., Watne, Å. K., Hammes, J., Lutz, A., Petäjä, T., Hallquist, M., Bilde, M., and Glasius, M.: High-Molecular Weight Dimer Esters Are Major Products in Aerosols from α-Pinene Ozonolysis and the Boreal Forest, Environmental Science & Technology Letters, 3, 280-285, 10.1021/acs.estlett.6b00152, 2016.
- Kurtén, T., Tiusanen, K., Roldin, P., Rissanen, M., Luy, J.-N., Boy, M., Ehn, M., and Donahue, N.: α-Pinene Autoxidation Products May
  Not Have Extremely Low Saturation Vapor Pressures Despite High O:C Ratios, The Journal of Physical Chemistry A, 120, 25692582, 10.1021/acs.jpca.6b02196, 2016.
  - Lee, B. H., Mohr, C., Lopez-Hilfiker, F. D., Lutz, A., Hallquist, M., Lee, L., Romer, P., Cohen, R. C., Iyer, S., Kurtén, T., Hu, W., Day, D. A., Campuzano-Jost, P., Jimenez, J. L., Xu, L., Ng, N. L., Guo, H., Weber, R. J., Wild, R. J., Brown, S. S., Koss, A., de Gouw, J., Olson, K., Goldstein, A. H., Seco, R., Kim, S., McAvey, K., Shepson, P. B., Starn, T., Baumann, K., Edgerton, E. S., Liu, J., Shilling, J. E., Miller, D. O., Brune, W., Schobesberger, S., D'Ambro, E. L., and Thornton, J. A.: Highly functionalized organic nitrates in the southeast United States: Contribution to secondary organic aerosol and reactive nitrogen budgets, Proceedings of the National Academy of Sciences, 113, 1516-1521, 10.1073/pnas.1508108113, 2016.
  - McFiggans, G., Mentel, T. F., Wildt, J., Pullinen, I., Kang, S., Kleist, E., Schmitt, S., Springer, M., Tillmann, R., Wu, C., Zhao, D., Hallquist, M., Faxon, C., Le Breton, M., Hallquist, A. M., Simpson, D., Bergstrom, R., Jenkin, M. E., Ehn, M., Thornton, J. A., Alfarra, M. R., Bannan, T. J., Percival, C. J., Priestley, M., Topping, D., and Kiendler-Scharr, A.: Secondary organic aerosol reduced by mixture of atmospheric vapours, Nature, 565, 587-593, 10.1038/s41586-018-0871-y, 2019.
  - McGraw, R., and Zhang, R.: Multivariate analysis of homogeneous nucleation rate measurements. Nucleation in the p-toluic acid/sulfuric acid/water system, The Journal of chemical physics, 128, 064508, 10.1063/1.2830030, 2008.
- Mentel, T. F., Springer, M., Ehn, M., Kleist, E., Pullinen, I., Kurtén, T., Rissanen, M., Wahner, A., and Wildt, J.: Formation of highly oxidized multifunctional compounds: autoxidation of peroxy radicals formed in the ozonolysis of alkenes deduced from structure–product relationships, Atmos Chem Phys, 15, 6745-6765, 10.5194/acp-15-6745-2015, 2015.
  - Mohr, C., Lopez-Hilfiker, F. D., Yli-Juuti, T., Heitto, A., Lutz, A., Hallquist, M., D'Ambro, E. L., Rissanen, M. P., Hao, L. Q., Schobesberger, S., Kulmala, M., Mauldin, R. L., Makkonen, U., Sipila, M., Petaja, T., and Thornton, J. A.: Ambient observations of dimers from terpene oxidation in the gas phase: Implications for new particle formation and growth, Geophys Res Lett, 44, 2958-2966, 10.1002/2017gl072718, 2017.
  - Molteni, U., Bianchi, F., Klein, F., El Haddad, I., Frege, C., Rossi, M. J., Dommen, J., and Baltensperger, U.: Formation of highly oxygenated organic molecules from aromatic compounds, Atmospheric Chemistry and Physics, 18, 1909-1921, 10.5194/acp-18-1909-2018, 2018.
- Paulsen, D., Dommen, J., Kalberer, M., Prevot, A. S., Richter, R., Sax, M., Steinbacher, M., Weingartner, E., and Baltensperger, U.: Secondary organic aerosol formation by irradiation of 1,3,5-trimethylbenzene-NOx-H2O in a new reaction chamber for atmospheric chemistry and physics, Environmental science & technology, 39, 2668-2678, 10.1021/es0489137, 2005.

Manuscript under review for journal Atmos. Chem. Phys.

Discussion started: 17 May 2019

© Author(s) 2019. CC BY 4.0 License.



5



- Rodigast, M., Mutzel, A., and Herrmann, H.: A quantification method for heat-decomposable methylglyoxal oligomers and its application on 1,3,5-trimethylbenzene SOA, Atmospheric Chemistry and Physics, 17, 3929-3943, 10.5194/acp-17-3929-2017, 2017.
- Sato, K., Takami, A., Kato, Y., Seta, T., Fujitani, Y., Hikida, T., Shimono, A., and Imamura, T.: AMS and LC/MS analyses of SOA from the photooxidation of benzene and 1,3,5-trimethylbenzene in the presence of NO, Atmos Chem Phys, 12, 4667-4682, 10.5194/acp-12-4667-2012, 2018.
- Shrivastava, M., Cappa, C. D., Fan, J. W., Goldstein, A. H., Guenther, A. B., Jimenez, J. L., Kuang, C., Laskin, A., Martin, S. T., Ng, N. L., Petaja, T., Pierce, J. R., Rasch, P. J., Roldin, P., Seinfeld, J. H., Shilling, J., Smith, J. N., Thornton, J. A., Volkamer, R., Wang, J., Worsnop, D. R., Zaveri, R. A., Zelenyuk, A., and Zhang, Q.: Recent advances in understanding secondary organic aerosol: Implications for global climate forcing, Rev. Geophys., 55, 509-559, 10.1002/2016rg000540, 2017.
- Trostl, J., Chuang, W. K., Gordon, H., Heinritzi, M., Yan, C., Molteni, U., Ahlm, L., Frege, C., Bianchi, F., Wagner, R., Simon, M., Lehtipalo, K., Williamson, C., Craven, J. S., Duplissy, J., Adamov, A., Almeida, J., Bernhammer, A. K., Breitenlechner, M., Brilke, S., Dias, A., Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Gysel, M., Hansel, A., Hoyle, C. R., Jokinen, T., Junninen, H., Kangasluoma, J., Keskinen, H., Kim, J., Krapf, M., Kurten, A., Laaksonen, A., Lawler, M., Leiminger, M., Mathot, S., Mohler, O., Nieminen, T., Onnela, A., Petaja, T., Piel, F. M., Miettinen, P., Rissanen, M. P., Rondo, L., Sarnela, N., Schobesberger, S., Sengupta, K., Sipilaa, M., Smith, J. N., Steiner, G., Tome, A., Virtanen, A., Wagner, A. C., Weingartner, E., Wimmer, D., Winkler, P. M., Ye, P. L., Carslaw, K. S., Curtius, J., Dommen, J., Kirkby, J., Kulmala, M., Riipinen, I., Worsnop, D. R., Donahue, N. M., and Baltensperger, U.: The role of low-volatility organic compounds in initial particle growth in the atmosphere, Nature, 533, 527-+, 10.1038/nature18271, 2016.
- Vereecken, L., and Peeters, J.: A structure–activity relationship for the rate coefficient of H-migration in substituted alkoxy radicals, 10.1039/C0CP00387E, 2010.
  - Wang, S., Wu, R., Berndt, T., Ehn, M., and Wang, L.: Formation of Highly Oxidized Radicals and Multifunctional Products from the Atmospheric Oxidation of Alkylbenzenes, Environmental science & technology, 51, 8442-8449, 10.1021/acs.est.7b02374, 2017.
  - Watne, Å. K., Psichoudaki, M., Ljungström, E., Le Breton, M., Hallquist, M., Jerksjö, M., Fallgren, H., Jutterström, S., and Hallquist, Å. M.: Fresh and Oxidized Emissions from In-Use Transit Buses Running on Diesel, Biodiesel, and CNG, Environmental science & technology, 52, 7720-7728, 10.1021/acs.est.8b01394, 2018.
  - WHO: Ambient air pollution: A global assessment of exposure and burden of disease, 2016.
  - Wyche, K. P., Monks, P. S., Ellis, A. M., Cordell, R. L., Parker, A. E., Whyte, C., Metzger, A., Dommen, J., Duplissy, J., Prevot, A. S. H., Baltensperger, U., Rickard, A. R., and Wulfert, F.: Gas phase precursors to anthropogenic secondary organic aerosol: detailed observations of 1,3,5-trimethylbenzene photooxidation, Atmos Chem Phys, 9, 635-665, 10.5194/acp-9-635-2009, 2009.
- Yan, C., Nie, W., Aijala, M., Rissanen, M. P., Canagaratna, M. R., Massoli, P., Junninen, H., Jokinen, T., Sarnela, N., Hame, S. A. K., Schobesberger, S., Canonaco, F., Yao, L., Prevot, A. S. H., Petaja, T., Kulmala, M., Sipila, M., Worsnop, D. R., and Ehn, M.: Source characterization of highly oxidized multifunctional compounds in a boreal forest environment using positive matrix factorization, Atmos Chem Phys, 16, 12715-12731, 10.5194/acp-16-12715-2016, 2016.