



Impacts of an Intense Wildfire Smoke Episode on Surface Radiation, Energy and Carbon Fluxes in Southwestern British Columbia, Canada

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Abstract

A short, but severe, wildfire smoke episode in July 2015, with an aerosol optical depth (AOD) approaching nine, had a significant impact on air quality, radiation and energy
20 budgets across four land use types, and elicited a clear ecosystem response with respect to carbon fluxes at a bog and a forested site. Greatest impacts on radiation and energy budgets were observed at the forested site where the role of canopy architecture, and the complex physiological responses to an increase in diffuse radiation were most important. With widespread standing water, and little physiological control on evapotranspiration,
25 the impacts on the partitioning of turbulent fluxes were modest at the bog compared to the physiologically dominated fluxes at the forested site. Despite the short duration and singular nature of the event, there was evidence of a diffuse radiation fertilization effect when AOD was less than two. With lighter smoke, both the wetland and forested site appeared to show enhanced photosynthetic activity (a greater sink for carbon-dioxide).
30 However, with dense smoke the forested site became a strong source. Given the extensive forest cover in the Pacific Northwest and the growing importance of forest fires in the region, these results suggest that wildfire aerosol during the growing season potentially



plays an important role in the regional ecosystem response to smoke and ultimately the carbon budget of the region.

1. Introduction

5 Wildfire activity is projected to increase in frequency and duration over the next century in western North America, primarily as a result of increased summer temperatures, persistent drought and reduced snowpack accompanying climate change (IPCC, 2014; Setelle et al. 2014). In addition to the obvious impacts on visibility and air quality, aerosols arising from biomass burning scatter and absorb solar radiation (direct effect)
10 while also influencing cloud processes by acting as cloud condensation nuclei (indirect effect) (IPCC, 2014). Furthermore, and of particular focus in this work, recent studies point to significant impacts on surface turbulent and radiative fluxes, boundary layer stability and energetics (including cloud development) and carbon exchange between biosphere and the atmosphere (Li et al., 2017).

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Radiative impacts of biomass burning are well documented in a variety of settings including South America (Moreira et al., 2017, Sena et al. 2013; Schafer et. al 2002), Africa (Schafer et. al 2002), Spain (Calvo et al. 2010), Russia (Chubarova et al. 2012, Pere et al. 2014), North America (Markowicz et. al. 2017, Vant-Hull et al. 2005) and Asia
20 (Wang et al. 2007). However, there is a growing literature concerned with impacts of smoke plumes on atmospheric boundary layer dynamics as well as surface radiation and energy budgets. For example, Taubman et al. (2004) investigated the impact of a wildfire plume underlain by an urban haze layer in Virginia and Maryland, USA, when aerosol optical depth (AOD) at 500 nm (τ_{500}) varied between 0.42 ± 0.06 and 1.53 ± 0.21 . In that
25 case, atmospheric absorption of solar radiation by the smoke and haze layers was nearly equal to the attenuation at the surface and resulted in net cooling at the surface and heating of the air aloft, thereby increasing stability. Absorption of solar radiation in the dense smoke layer maintained a morning subsidence inversion and thereby created a positive feedback loop preventing vertical mixing and dilution of the smoke plume itself.
30 Wang and Christopher (2006) report a similar broad range of impacts on surface radiation/energy budgets, and boundary layer dynamics in a modeling study of the impact of Central American Biomass Burning on source region as well as the Southeastern US



(for $\tau_{550}=0.09$). At the plant canopy scale, Yamosoe et al. (2006) have focused on the impact of biomass burning aerosol on Amazonian forests and have noted an increase in diffuse radiation within the canopy combined with a reduction in total photosynthetically active radiation (PAR) at the top of the canopy. These impacts affected sensible and latent fluxes as well as net ecosystem exchange (NEE) of carbon dioxide (CO_2). Subsequently, Steiner et al. (2013) have explored such ecosystem responses using data from six US FLUXNET sites and demonstrate that high AOD reduces midday net radiation by 6%–65% coupled with a 9%–30% decrease in sensible and latent heat fluxes. Niyogi et al. (2004), in an examination of six AmeriFlux sites, conclude that aerosols can exert a significant impact on net CO_2 exchange (perhaps more so than clouds) whereby the CO_2 sink is increased with aerosol loading for forest and croplands. This effect has become known as the diffuse radiation fertilization effect (DRF) whereby an increase in photosynthesis results from a trade-off between decreased solar radiation and increased light scattering during clouds or smoke (Park et al. 2017, and references therein). It is suggested that the magnitude of the effect is controlled by canopy architecture, leaf area index and plant functional type.

In western Canada, previous studies have examined the chemistry and transport of smoke plumes, as well as the impacts on local air quality (Cottle et al. 2014, McKendry et al. 2011; McKendry et al. 2011). However to date, the impacts on radiation and energy budgets, boundary layer dynamics and ecosystems have not been addressed for biomass burning associated with the coastal temperate coniferous forest biome. As the risk of wildfire increases in such areas (Setelle et al., 2014) biomass burning is likely to have non-trivial impacts on surface climates as well as ecosystem productivity and the carbon cycle in Canada's most productive ecozone.

During the summer of 2015, a rare opportunity arose to investigate such impacts under clear skies and for unusually high AOD ($\tau_{500} \sim 2.5$). A period of prolonged drought and elevated temperatures resulted in sustained wildfire activity throughout the Pacific Northwest. In early July 2015, a particularly intense event had significant impacts on air quality and visibility in southwestern British Columbia. Ground level hourly $\text{PM}_{2.5}$ (particulate matter less than $2.5 \mu\text{m}$ in diameter) concentrations approached $200 \mu\text{g m}^{-3}$ in



the city of Vancouver on 5 July and were associated with smoke emanating from approximately 150 km to the north-east in the Pemberton region. (Figure 1). By 7 July reports noted that these Elaho, Boulder Creek and Nahatlatch fires had spread to a combined area of approximately 30,000 ha.

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In this study, we focus on the impacts of relatively “fresh” smoke (~1-2 days old) from these intense temperate coniferous forest fires on the radiation budgets, turbulent fluxes and ecosystem responses of four distinct land-use types (a wetland, an urban residential area, an agricultural grass field, and a coniferous forest). In so doing we add a new geographic setting to the growing catalog of such ecosystem impacts, and compare the results with studies from other regions. Furthermore, the availability of sunphotometer and aerosol LiDAR data from the immediate area, greatly enriches the information available for interpretation of this event.

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15 **2. Background and Data Sources**

2.1 Synoptic Overview

During 2-6 July 2015, western Canada was under the influence of a 500 hPa ridge of high pressure centred off-shore at 135° W. This resulted in northwesterly upper level flow across southwestern British Columbia. At the surface, a thermal trough was located along the western Cordillera, a pattern associated with poor air quality in the region (M^cKendry, 1994). Vancouver International Airport recorded maximum daily temperatures in the range 25-27°C from 2-6 July with nighttime minima of 21°C. Skies were generally cloudless with no precipitation recorded and maximum wind gusts were of ~10 m s⁻¹.

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Commercial aircraft soundings (Aircraft Meteorological DATA Relay - AMDAR) from Vancouver International Airport (YVR) departures and arrivals show development of a strong surface based inversion late on 4 July that persisted through 5 July 2015, and coinciding with smoke arrival mid-afternoon on 5 July over the western edge of the Lower Fraser Valley (see supplementary material). At that time, the inversion top was at ~500 m above ground and was ~10 K in magnitude. As shown below, this strong inversion effectively trapped the smoke plume below it and was likely responsible for the high particulate matter concentrations and poor visibility observed on 5 July. By 6

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July this capping inversion was no longer present and likely disappeared as a result of the evolving weather pattern and advection.

2.2 CORALNet (Canadian Operational Research Aerosol Lidar Network)

5 CORALNet is a semi-autonomous network of ground-based LiDARs, one of which has operated since 2008 at the University of British Columbia (UBC – Totem Field – see Figure 1). This remotely controlled facility was housed in a cargo trailer with modifications including a roof hatch assembly, basic meteorological tower, radar interlock system, climate control system and levelling stabilizers. A Continuum Inlite III
10 (small footprint) laser operating at 1064/532 nm simultaneously with a pulse repetition rate of 10 Hz is the foundation of the system. The upward-pointing system measures the return signal in three channels (1064 nm, and two polarization channels at 532 nm). The system is described in detail by Strawbridge (2013), and an example of its application shown in Cottle et al. (2014).

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2.3 AERONET/AEROCAN

The global AERONET (Aerosol RObotic NETwork) has operated since 1993 and is focused on measurements of vertically integrated aerosol properties using the CIMEL sunphotometer/sky radiometer instrument (Holben et al., 1998). AEROCAN CIMELs
20 (AEROCAN is the Canadian sub-network of AERONET) include a facility on Saturna Island 55 km to the south of the UBC CORAL-net site. Here, solar irradiances are acquired across eight spectral channels (340, 380, 440, 500, 670, 870, 1020 and 1640 nm) that are transformed into three processing levels of aerosol optical depth (AOD); 1.0 – non-cloud screened; 1.5 – cloud screened; and 2.0 – cloud screened and quality assured.
25 McKendry et al. (2011) demonstrated the application of these data to the transport of California wildfire plumes. In this paper, as in McKendry et al. (2011), the SDA algorithm was applied to Level 1.0 AOD spectra in order to better delineate the strongly varying contribution of fine mode smoke particles at a reference wavelength of 500 nm. Level 1.0 input AODs were chosen to minimize “false negative” smoke-AOD rejection
30 attributed to the Level 1.5 cloud-screening algorithm.



2.4 Radiative and Turbulent Flux Data

The smoke event of July 2015 coincided with a period in which routine long term measurements of surface radiation and turbulent fluxes (sensible and latent heat using the eddy-covariance method) were made at three sites in the region, while the at a fourth site
5 only the radiation budget was observed (Table 1). Photographs of the sites are shown in the supplementary materials.

Buckley Bay (Ca-Ca3) is a flux tower with eddy-covariance and radiation sensors measuring exchange between a coniferous forest stand (Douglas-fir, 27 years old) and the
10 atmosphere. The site is located on the eastern slopes of the Vancouver Island Range, about 150 km to the west of Vancouver. Full descriptions of the site and the instruments can be found in Humphreys et al. (2006) and Chen et al. (2009). Burns Bog (Ca-DBB) is a floating platform with eddy-covariance and radiation instrumentation on an open wetland with mosses, sedges, and a significant fraction of standing water. Further details
15 of the site are described in Christen et al. (2016) and Lee et al. (2017). Vancouver-Sunset (Ca-VSu) is an urban observational tower above a residential detached urban neighborhood. Details of the instrumentation can be found in Crawford et al. (2014). Vancouver-UBC is a climate station on the Campus of the University of British Columbia that features a full set of radiation measurements.

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Table 1: Measurements and site characteristics

Site	Fluxnet ID	Energy Budget	Radiation Budget	NEE	Coordinates (WGS-84)
Buckley Bay Coniferous Forest	Ca-Ca3	•	•	•	124° 54' 1.44" W 49° 32' 4.63" N
Burns Bog Wetland	Ca-DBB	•	•	•	122° 59' 5.60" W 49° 07' 45.59" N
Vancouver-UBC Grass	-		•		123° 14' 56.41" W 49° 15' 19.50" N
Vancouver-Sunset Residential Urban	Ca-VSu	•	•		123° 4' 42.24" W 49° 13' 33.96" N



Based on descriptions and conventions described in Oke (1987), the surface radiation budget can be defined as:

$$Q^* = K_{\downarrow} - K_{\uparrow} + L_{\downarrow} - L_{\uparrow}$$

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where Q^* is the net all-wave radiation), K_{\downarrow} is the shortwave irradiance comprising direct and diffuse solar radiation, K_{\uparrow} is the reflected shortwave radiation, L_{\downarrow} is the longwave (“thermal”) irradiance from the sky and L_{\uparrow} the longwave radiation emitted and reflected from the surface (all in W m^{-2}).

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The ratio $K_{\uparrow}/K_{\downarrow}$ is the surface albedo (α) and is the shortwave reflectance of the surface in the solar band. K_{ext} is the extra-terrestrial solar radiation and represents the flux density of solar radiation falling at the outer edge of atmosphere and is computed based on date, time and latitude at the site. The ratio of K_{\downarrow}/K_{ext} is a measure of the bulk transmissivity of the atmosphere to shortwave radiation. Photosynthetically Active Radiation (PAR), measured at Burns Bog only, is shortwave radiation in the range 440-670 nm and is typically expressed in terms of photon flux density ($\mu\text{mol m}^{-2} \text{s}^{-1}$).

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Furthermore, the non-radiative partitioning of energy partitioning at the surface can be defined using the surface energy balance (Oke, 1987):

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$$Q^* = Q_H + Q_E + Q_G$$

where Q_H is the turbulent (convective) sensible heat flux to the atmosphere, Q_E is the turbulent (convective) latent heat exchange with the atmosphere (including evaporation and transpiration) and Q_G is the conductive exchange of energy with the underlying substrate (all in W m^{-2}). The Bowen ratio is defined as $\beta = Q_H/Q_E$ and is a measure of the partitioning of the turbulent heat fluxes. β is dependent on availability of water at the surface as well plant physiology and has important consequences for surface climates by influencing both surface temperature and humidity.

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Net ecosystem exchange (NEE in $\mu\text{mol m}^{-2} \text{s}^{-1}$) is used in quantifying the carbon balance of an ecosystem and is ecosystem respiration (R_e) minus gross ecosystem photosynthesis (GEP), i.e., $\text{NEE} = R_e - \text{GEP}$. NEE is negative when the ecosystem is acting as a CO_2 sink, and positive when it is acting as a CO_2 source.

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3. Results

3.1. Satellite, Lidar and Sunphotometer Observations

MODIS (Moderate Resolution Imaging Spectrometer) imagery for the period is shown in Figure 2a-d. On 4 July 2015 the region was cloud and smoke free. By 5 July a plume of smoke from the fires in the Elaho valley near Pemberton is evident and extends across the southern and central portion of Vancouver Island (including Buckley Bay, but not the three mainland sites). At this time, a “wall of smoke” extended broadly from northwest to southeast along the Strait of Georgia and slightly to the west of the city of Vancouver. This smoke moved across the city of Vancouver at approximately 15:00 Pacific Daylight Time (PDT) on 5 July 2015 (photographic evidence is shown in supplementary material). HYSPLIT (Hybrid Single Particle Lagrangian Integrated Trajectory Model) modeling (see supplementary information) at this time confirmed the source, shape and extent of the plume. By 6 July, the plume had dispersed eastward and was accompanied by cloud to the west of Vancouver Island with a signature consistent with a coastally trapped disturbance (Reason and Dunkley, 1993) or marine “stratus surge” as it is commonly known in the region. However, the region to the east of Vancouver Island remained cloud free, and remained so during the 6 July when dense smoke was still evident across southwestern British Columbia including all four of the measurement sites.

The impact of smoke on air quality in the vicinity of Vancouver is shown in Figure 3. LiDAR imagery (Figure 3c) shows an elevated layer of smoke over the region at $\sim 2000\text{m}$ elevation prior to the arrival of a “wall of smoke” at ground level at approximately 15:00 PDT on 5 July (depicted by the vertical dashed line). Ground level smoke remained in a shallow layer until approximately 6:00 PDT on 6 July. Subsequently, smoke continued to persist over the region but was confined to a shallow layer at $\sim 1750\text{m}$ elevation AGL. Smoke again descended toward ground level on 7 July but did not reach the surface. Consequently, PM concentrations at Vancouver International Airport (Figure 3a) peaked



(reaching $250 \mu\text{g m}^{-3}$) when smoke was at ground level between 15:00 on 5 July and 6:00 PDT on 6 July. Over the entire period there was a modest decrease in daytime maximum and minimum temperatures at Vancouver International Airport (Figure 3a). Further discussion on the role of smoke on temperatures is provided below.

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Analysis of sunphotometer (Saturna Island), LiDAR (UBC) MODIS, AQUA and CALIOP (Cloud-Aerosol Lidar with Orthogonal Polarization) together reveal a complex three-dimensional structure associated with the smoke event (with layers extending into the 3-6 km range AGL). The event, as shown in both LiDAR and CALIOP data consisted of multiple layers with the predominately fine mode particle signature of smoke confirmed in the AERONET (Saturna Island) data (Figure 3b). Prior to the arrival of the ground level smoke over Vancouver, fine mode AOD values at Saturna Island were very high (~ 9 at 11:00 PDT) and were (on the basis of detailed analysis of the MODIS and AQUA imagery) likely associated with the higher altitude (2-3 km) smoke plumes. The 15:00 PDT “wall of smoke” mentioned above is seen as a sharp fine mode AOD rise at Saturna following the decay of the strong 11:00 peak of the 2-3 km layer (a rise that started around 12:30 PDT : the $2 \frac{1}{2}$ hour difference being a function of the Saturna to UBC transport time and the time that a significant increase in fine mode AOD could be detected at Saturna).

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3.2 Impact on Radiation and Energy Budgets

The course of diurnal radiation budget components at each site is shown in Figure 4, while daily averages are listed in Table 2. On both 3 and 4 July, all sites show a smooth diurnal course of radiation components consistent with summer clear sky conditions. On these days, mean daily atmospheric bulk transmissivity (K_{\downarrow}/K_{ext}) was approximately 80% at all sites (Table 2). The most dramatic impact of the smoke plume on radiation components occurred on July 5 at Buckley Bay when the mean daily transmissivity dropped to 40% with a reduction in midday K_{\downarrow} of 49% (to 475 W m^{-2}) compared to midday values on 3 and 4 July ($\sim 920 \text{ W m}^{-2}$). This is consistent with satellite imagery (Figure 2) which shows the smoke layer persisting over Buckley Bay for the entire day on 5 July. At mainland sites, the late arrival of smoke at approximately 15:00 PDT on 5 July is evident in the late afternoon K_{\downarrow} but had less impact on daily totals. Instead, at these

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mainland sites the biggest impact of the smoke occurred on 6 July when daily transmissivities dropped to 52-57% and peak midday K_{\downarrow} values were reduced by approximately 15-25% compared with those observed on the 3 and 4 July. On this day, LiDAR imagery shows the smoke layer to be at a higher elevation with less intense backscatter than seen late on 5 July (Figure 4a).

Table 2: Daily averages of radiative, turbulent fluxes, Bowen ratio and carbon dioxide fluxes 3-7 July for each of the four sites

	3 July 2015	4 July 2015	5 July 2015	6 July 2015	7 July 2015
K_{ext}	MJ m ⁻² day ⁻¹				
	37.0	36.8	36.7	36.4	36.2
Shortwave K_{\downarrow}	MJ m ⁻² day ⁻¹				
Burns Bog	29.5 (78%)*	29.4 (78%)	25.3 (69%)	20.8 (57%)	23.8 (65%)
Van.-Sunset	28.3 (76%)	28.1 (76%)	23.4 (63%)	20.8 (57%)	21.6 (60%)
Van.- UBC	28.2 (76%)	27.8 (76%)	21.5 (59%)	18.9 (52%)	19.7 (54%)
Buckley Bay	29.8 (81%)	29.9 (81%)	14.4 (39%)	21.2 (58%)	24.4 (67%)
Albedo (α)					
$K_{\uparrow}/K_{\downarrow}$					
Burns Bog	0.18	0.18	0.18	0.18	0.17
Van.-Sunset	0.16	0.15	0.17	0.17	0.16
Van.-UBC	0.29	0.30	0.37	0.35	0.34
Buckley Bay	0.13	0.13	0.17	0.15	0.14
Sensible Heat Q_H	DAYTIME (Wm ⁻²)				
Burns Bog	67.3	63.4	30.5	60.0	97.6
Van.-Sunset	250.7	247.3	136.5	130.1	176.4
Buckley Bay	232.8	217.5	41.4	136.9	179.3
Latent Heat Q_E	DAYTIME (Wm ⁻²)				
Burns-Bog	116.8 ($\beta=0.58$)	112.8 ($\beta=0.56$)	99.6 ($\beta=0.31$)	89.8 ($\beta=0.67$)	96.4 ($\beta=1.01$)
Van.-Sunset	53.6 ($\beta=4.68$)	57.8 ($\beta=4.28$)	62.7 ($\beta=2.18$)	48.3 ($\beta=2.69$)	39.0 ($\beta=4.52$)
Buckley Bay	72.4 ($\beta=3.21$)	70.1 ($\beta=3.10$)	49.1 ($\beta=0.84$)	60.3 ($\beta=2.27$)	53.5 ($\beta=3.35$)



CO ₂ -NEE (daily mean)	g C m ⁻² day ⁻¹				
Burns Bog	-1.67	-1.64	-2.47	-3.64	-4.24
Buckley Bay	0.16	1.26	1.13	-1.35	-2.31

* percentage of K_{ext} (extraterrestrial radiation)

Daytime Bowen ratio $\beta = Q_H/Q_E$

- 5 With respect to remaining radiation budget components, arrival of smoke at all sites was marked by a reduction in Q^* . However, variations in L_{\downarrow} and L_{\uparrow} were subtle and point to only modest changes in surface or atmospheric temperatures (Figure 3a). Albedo (Table 2) also increased at the two Vancouver sites as well as the Buckley Bay site with the arrival of smoke and likely is a consequence of the reduction in specular reflection during
- 10 direct solar irradiance and an increase in diffuse reflection.

Diurnal impacts of the smoke event on atmospheric transmissivity are shown in Figure 5 where a clear non-smoke day (3 July 2015) is directly compared with 6 July 2015. Of note, the impact of the low level smoke is apparent in the significant reductions

15 in transmissivity throughout the day. However, near sunrise and sunset, 3 and 6 July have almost the same irradiance at low sun angle, presumably due to the presence of more diffuse light. Secondly, the impacts are the same across all sites/ecosystems and therefore demonstrate a clear regional signal consistent with the widespread smoke distribution shown in Figure 2c.

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The course of sensible (Q_H) and latent (Q_E) turbulent heat fluxes are shown in Figure 6 and summarized in Table 2. As with K_{\downarrow} above, the most significant impact on Q_H was at Buckley Bay on 5 July where it decreased to 18% (i.e., 41.4 W m⁻²) of clear sky mean daytime time values. At Vancouver-Sunset and Burns Bog, the greatest reductions

25 occurred on 6th July and were 52% and 90% respectively compared to the daytime values on the clear-sky days preceding the episode. The impacts on Q_E were less than for Q_H at all sites. At Burns Bog and Vancouver-Sunset, minima for Q_E and Q_H occurred on 6 July whereas at Buckley Bay the minimum occurred on 5 July. At all sites, β was reduced on



5 and 6 July with the greatest reduction at Buckley Bay on 5 July. The latter was the result of the large reduction in Q_H at that site (82%) and the relatively small reduction in Q_E (32%). The switch from high direct radiation on 3 and 4 July to predominately diffuse radiation on 5 July was likely responsible for the marked reduction in Q_H as a consequence of reduced heating of leaves in a highly coupled forest canopy (Brümmer et al 2012) In summary, evapotranspiration was maintained at all sites with the wettest sites (Burns Bog) showing the lowest response to the smoke plume. However, impacts on Q_H were greatest at the forested site (Buckley Bay).

10 **3.3 Impact on CO₂/NEE and PAR**

Daily values for NEE are shown in Table 2 for two sites (Burns Bog and Buckley Bay) where CO₂ fluxes were measured over active vegetation. Throughout the smoke period, Burns Bog remained a net carbon sink, and showed an increasingly negative trend in NEE (stronger sink) over the duration of the smoke episode. On 3 July, under clear sky conditions and pre-arrival of smoke the bog was a net CO₂ sink (-1.67 g C m⁻² day⁻¹). This was consistent over the seven previous precipitation free days (mean -1.69 g C m⁻² day⁻¹). The peak radiative impact of the smoke at Burns Bog occurred on 6 July (albeit somewhat lesser in magnitude than occurred at Buckley Bay) and was associated with a daily NEE of -3.64 g C m⁻² day⁻¹ (net sink). This effect was even more pronounced on the following day.

Conversely, at the Buckley Bay forested site, the pre-smoke arrival daily NEE on 3 July showed a CO₂ neutral situation (0.16 g C m⁻² day⁻¹). Again, as with the Burns Bog site, this was consistent with the seven previous precipitation-free days (mean NEE of -0.08 g C m⁻² day⁻¹), confirming that during such mid-summer conditions (clear, warm and precipitation-free), the Buckley Bay site at the daily scale is broadly CO₂ neutral. When the peak impact on incoming shortwave radiation was felt on the 5th of July (with significantly greater solar attenuation than occurred at the three mainland sites), NEE became more strongly positive (a greater atmospheric source of CO₂) and then became a strong net sink on 6 and 7 July (-1.35 and - 2.31 g C m⁻² day⁻¹ respectively) when the smoke had started to disperse.



Figure 7 shows the course of the PAR/K_{\downarrow} ratio before and during the smoke episode at Burns Bog. Under clear sky conditions, PAR is roughly a fraction of K_{\downarrow} , and for long-term, and all weather conditions, Tortini et al. (2017) found a value of $1.798 \pm 0.026 \mu\text{mol J}^{-1}$ for Burns Bog. This value is comparable to the mid-day values during the first three smoke-free days (July 2 to 4) of $1,789 \mu\text{mol J}^{-1}$. However, with the arrival of smoke, ratios were reduced significantly to $1.609 \mu\text{mol J}^{-1}$. This suggests that during heavy smoke, there are fewer PAR photons available for photosynthesis per energy received.

4. Discussion

Based on the observations described above, the presence of a dense layer of wildfire smoke near the surface resulted in significant perturbation of both the radiation and energy budgets over a range of surface types in southwestern British Columbia in early July 2015. The effect was most strongly felt at the forested Buckley Bay site on Vancouver Island.

The dramatic attenuation of incoming shortwave radiation described in section 3.2 is entirely consistent with published literature for forest fire plumes described elsewhere and for a similar range of AOT_{500} . Perhaps the best analog is the 2010 fires in central Russia described by Chubarova et al. (2012) and Péré, et al. (2014). Chubarova et al. (2012) report a 40% loss of shortwave irradiance at $AOT_{500} = 2.5$ (their Figure 10), a value consistent with the losses of 30-50% across our four sites (Table 2) in daytime mean fluxes in K_{\downarrow} . Interestingly, Chubarova et al. (2012) observed much greater losses of ~65% for UV radiation (300–380nm) and ~80% for erythemally-weighted irradiance. For the same events, Péré, et al. (2014) examined the shortwave aerosol direct radiative forcing and its feedback on air and atmospheric temperature over Moscow. For τ_{340} in the range 2-4, wildfire aerosol caused a significant reduction of surface shortwave radiation (up to $70\text{--}84 \text{ Wm}^{-2}$ in diurnal averages) which is again consistent with the $\sim 100 \text{ Wm}^{-2}$ reduction over background in diurnal averages of K_{\downarrow} at the four British Columbia sites.

Similar agreement is apparent when compared with observed reductions of total solar irradiance by forest fire smoke in the Brazilian Amazon and Zambian Savanna (Schafer et al. 2002) At the four sites examined here, total solar irradiance (Table 2) during the fire



event represented 50-70% of background values. This is in broad agreement with Brazilian sites (Alta Floresta and Abracos Hill) in Schafer et al. (2002; Figure 1a) which show a reduction of ~68% for K_{\downarrow} over background values for $\tau_{500} = 2.5$. In their study, the African grassland sites show impacts of similar magnitude at somewhat lower AOT values, a likely consequence of the different fuel type, combustion temperatures and aerosol optical properties of the aerosol generated in such fires.

As with radiation budget components, impacts on turbulent heat fluxes were variable across the four sites with the greatest impact at the forested Buckley Bay site where β was reduced significantly on 5 July to 0.84 from values of ~3.2 on the preceding clear days. Again, these results are broadly consistent with prior studies elsewhere showing that the impact of aerosol is to reduce K_{\downarrow} (but perhaps increase diffuse radiation) and hence Q^* , Q_H and Q_E , with the partitioning of the turbulent fluxes β appearing to be ecosystem dependent (Steiner et al, 2103). It is important to note however, that FLUXNET data cited by Steiner et al. (2013) for four forested sites, a grassland site and cropland site represent averages from quite different geographical settings than those considered here, and are for $\tau_{500} < 1.2$, significantly less than τ_{500} values observed in this case.

With respect to Buckley Bay observations, where canopy effects are most important, Yamasoe et al. (2006) offer perhaps a more germane comparison in the context of the Amazon rainforest and for smoke AOT's of a similar magnitude to those observed at Buckley Bay on 5 July 2015. In their study, both Q_H and Q_E were observed to decrease along with a decrease in photosynthetically active radiation (PAR) due to aerosol attenuation. In this study, Burns Bog offers an interesting contrast to the forested Buckley Bay site. With widespread standing water, and little physiological control on Q_E , the impacts on the partitioning of turbulent fluxes were modest compared to the physiologically dominated fluxes at Buckley Bay. The marked reduction in Q_H compared to Q_E (and resulting drop in β) at Buckley Bay clearly shows the dominating effect of canopy (stomatal) resistance over the much smaller aerodynamic resistance in this highly-coupled forest ecosystem (McNaughton and Jarvis 1983).



Despite the short duration and singular nature of the smoke event described here, this case study offers at least a tentative indication of the potential magnitude of a DRF effect in two quite different ecosystems in the Pacific Northwest. In both cases, the arrival of heavy smoke initiated a clear ecosystem response. Burns Bog, typically a CO₂ sink in clear summer conditions, became an even stronger sink with the arrival of smoke. Buckley Bay forest, generally CO₂ neutral in such conditions, became a strong source with the arrival of heavy smoke, and then returned to being a carbon sink on 6 and 7 July when the smoke had started to disperse. The latter hints that as the radiative impact of the smoke diminished, and AOT dropped below the critical threshold of two noted by Yamasoe et al. (2006) and Park et al. (2017), a delayed DRF effect may have been initiated that promoted photosynthesis within the canopy.

These broad patterns seem consistent with previous research in different environments (Yamasoe et al, 2006; Nyogi et al. 2004; Park et al., 2017). Yamasoe et al (2006), in the Amazon basin, show smoke caused an increase in the diffuse fraction of PAR, thereby enhancing transmission deeper into the canopy, leading to enhanced photosynthetic activity and CO₂ uptake for moderate τ_{500} . However, of particular relevance to this study, at high τ (>2), the magnitude of the CO₂ flux and NEE decreased, an effect they ascribed to low PAR values and potentially deleterious impacts of pollutants in the smoke itself. This effect has also been observed by Park et al. (2017) in the central Siberian taiga. There, when $\tau > 2$, a reduction of PAR and diffuse PAR occurred and the forest became a CO₂ source. The observation in this study that not only is PAR reduced in dense smoke, but also the ratio of PAR/K_{\downarrow} is diminished when compared to Tortini et al.'s (2017) typical values, also seems to be an potentially important factor contributing to the overall ecosystem response, and especially the magnitude of the DRF effect. We intend to explore this issue (spectral impacts of smoke) further in the context of more recent fire events and at Buckley Bay forested site in particular. It should be noted that the impact of smoke on the radiation budget at Burns Bog was significantly less than occurred at Buckley Bay (see table 2). It is therefore likely that the impact on AOT at this site was also diminished (and may not have exceeded $\tau = 2$) when compared with Buckley Bay. Our results, and those elsewhere, suggest that the ecosystem response to smoke is



dependent on the density of smoke and may well be highly variable spatially and temporally, and by ecosystem type. Clearly further research is required in western North America to identify the major drivers governing ecosystem response and also the impacts of longer term exposure to smoke.

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Finally, we were unable to quantitatively assess the impact of the smoke layer on atmospheric stability in this case. Elsewhere, it has been shown that dense smoke layers provide a positive feedback mechanism by increasing stability and inhibiting cloud formation. In the absence of a spatial array of vertical soundings and due to the rapidly
10 evolving synoptic situation (where advection was important) we were unable to quantify the radiative effects on the plume layer itself. From available aircraft AMDAR soundings, it was apparent that the plume was trapped by a strong inversion that preceded the arrival of the plume. Certainly, modeling studies in other settings suggest that similar smoke layers may be subject to radiative heating rates of $\sim 6 \text{ K day}^{-1}$ (Calvo et al. 2010,
15 Feingold et al. 2005, Stone et al. 2008) with significant cooling at the surface, thereby significantly enhancing stability. We propose that a modeling study would help elucidate the processes at play in this case.

5. Conclusions

20 The wildfire smoke episode of early July 2015 in southwestern British Columbia had a significant impact on air quality, radiation and energy budgets, and elicited a clear ecosystem response with respect to NEE of land ecosystems. Across the four land-use types monitored, impacts were variable, but consistent with published literature in other settings. The greatest impacts on radiation and energy budgets were observed at the
25 forested site where the role of canopy architecture, and the complex physiological responses to an increase in diffuse radiation were most important. Despite the short duration and singular nature of the event, there was evidence of a DRF effect when smoke density was lower than the threshold of $\tau=2$. With lighter smoke, both the wetland and forested site appeared to show enhanced photosynthetic activity (a greater carbon
30 sink). However, with dense smoke, and significantly reduced irradiance, the forested site became a strong source. This is consistent with literature suggesting that with dense smoke, within canopy PAR is reduced to a point where reduced photosynthetic activity



outweighs the DRF effect and the forest becomes a net carbon source (as at night). Given the extensive forest cover in the Pacific Northwest and the growing importance of forest fires in the region, these results suggest that wildfire aerosol potentially plays an important role in the regional ecosystem response to smoke and ultimately the carbon budget of the region. Due to the short duration of the event described here, we recommend further research, including modelling, to elucidate and generalize the patterns observed in this single case.

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Figure 1: Map with inset showing all places mentioned in text



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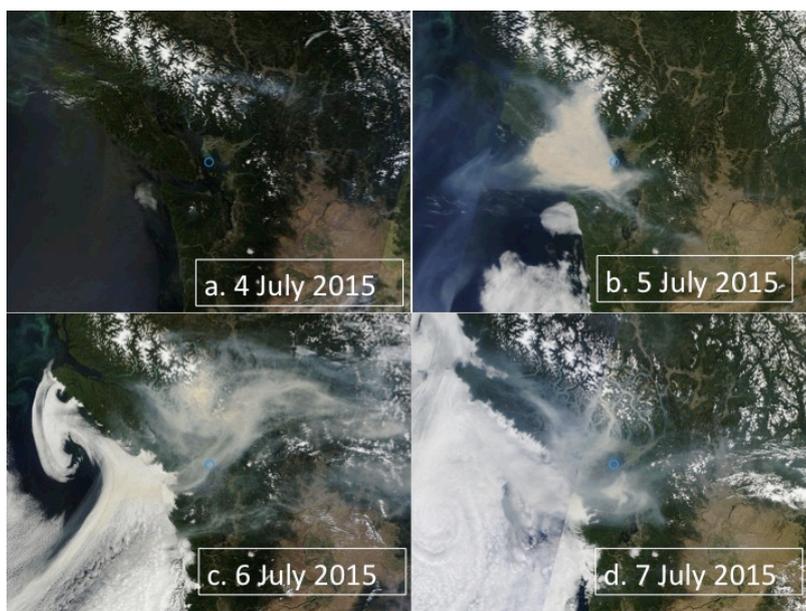
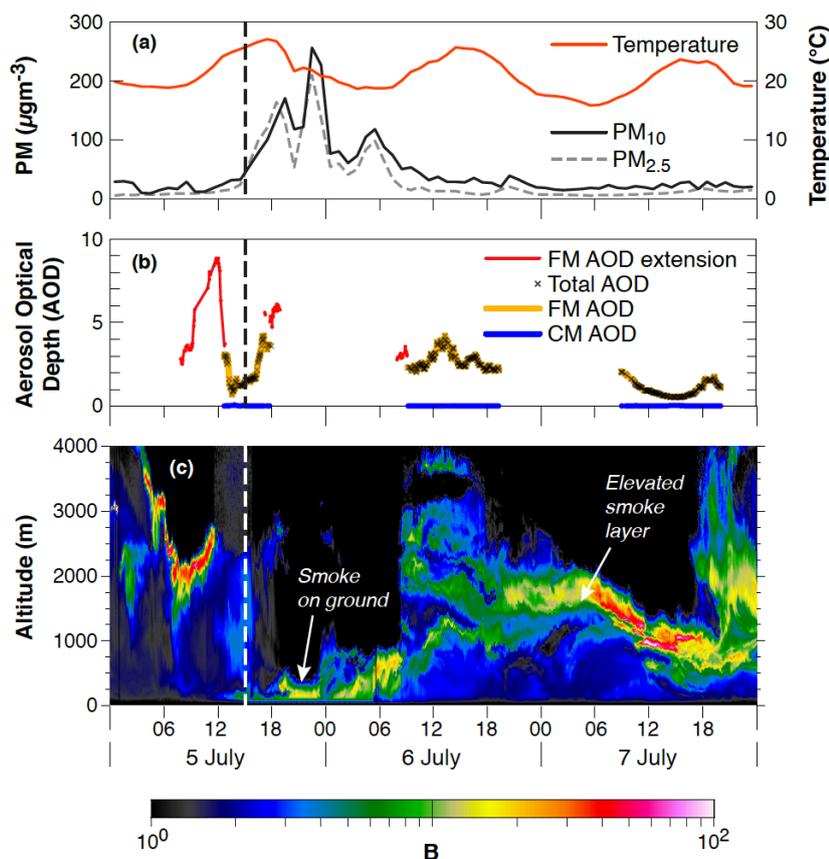


Figure 2: Modis Satellite Imagery for the period 4-7 July (Saturna Island shown in Blue circle)

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5 **Figure 3:** Time series for 5-7 July 2017 showing (a) PM_{2.5} and PM₁₀ observations, and temperature at Vancouver International Airport and (b) Aerosol Optical Depths for fine mode (FM) and coarse mode (CM) variation at Saturna Island. The “FM AOD extension” was obtained by assuming that the total AOD at the longer wavelengths of 675 and 870 nm was dominated by the fine mode AOD and extrapolating their AODs back to 500 nm
 10 (a choice necessitated by the fact that the extraordinarily large AODs at the shorter wavelengths were eliminated by AERONET processing) and (c) LiDAR backscatter from the UBC CORAL-NET site at 532 nm for the period. The red vertical line shows arrival of the low level “wall of smoke” around 3pm on 5 July.



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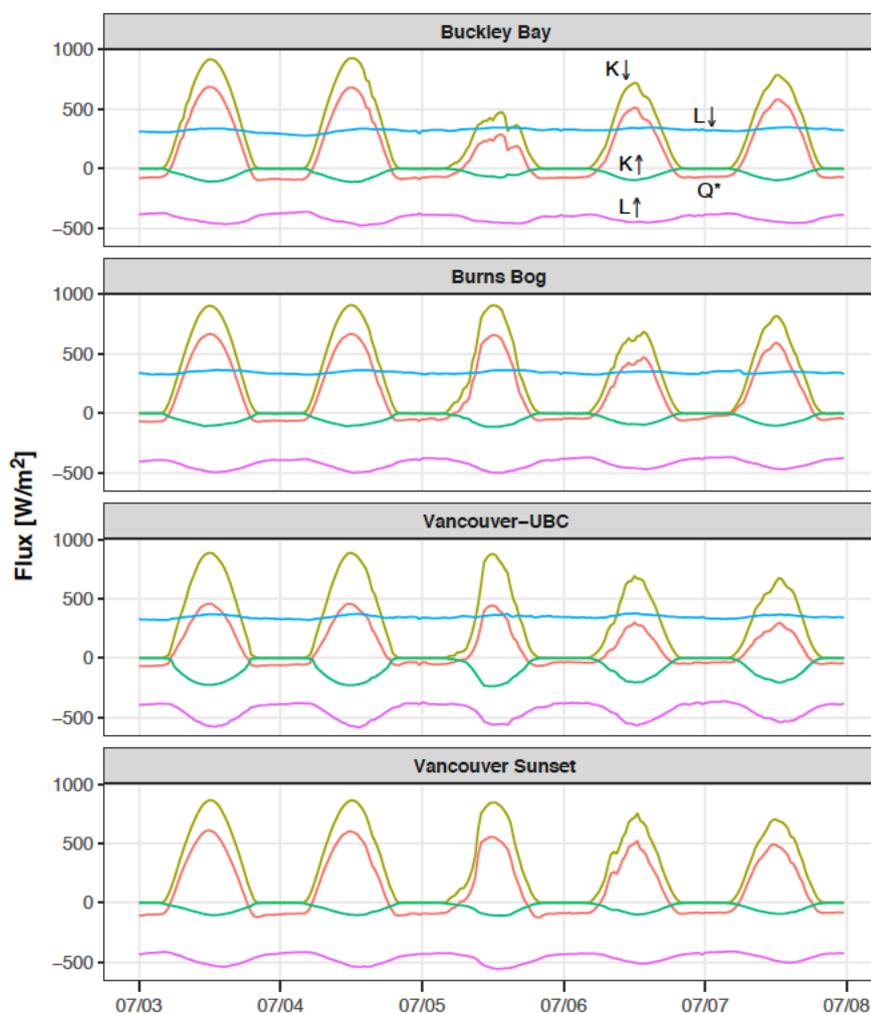


Figure 4: Radiation budget components using standard convention. Fluxes away from surface plotted as negative values.

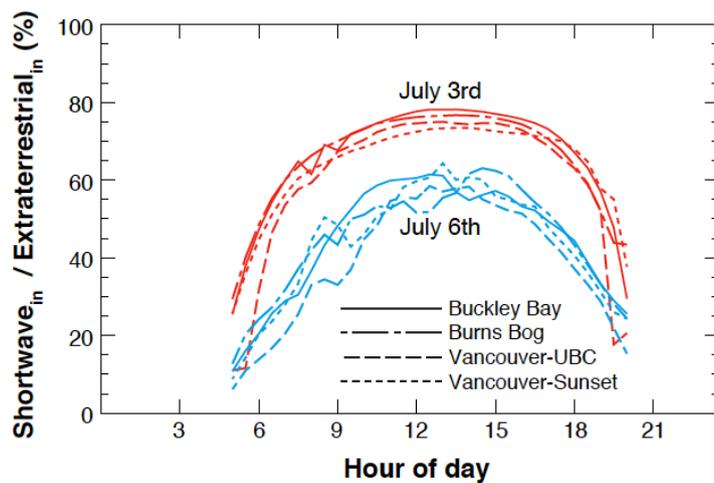


Figure 5: Diurnal impacts on incoming solar radiation at each site for 3 (cloudless day) and 6 (smoke) July.

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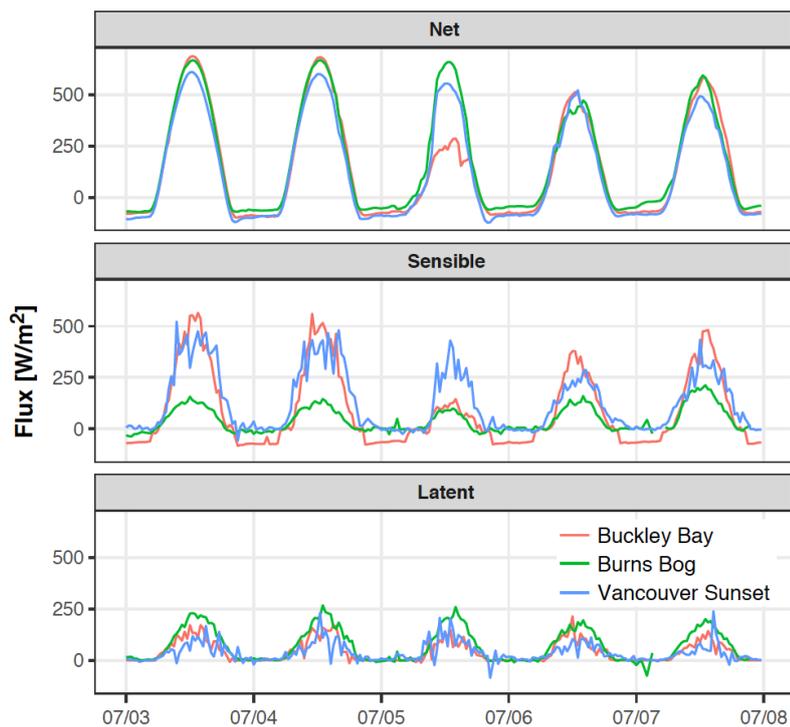


Figure 6: Net radiation (Q^*), sensible (Q_H) and latent (Q_E) heat fluxes at three sites. Fluxes away from surface are plotted as positive values.

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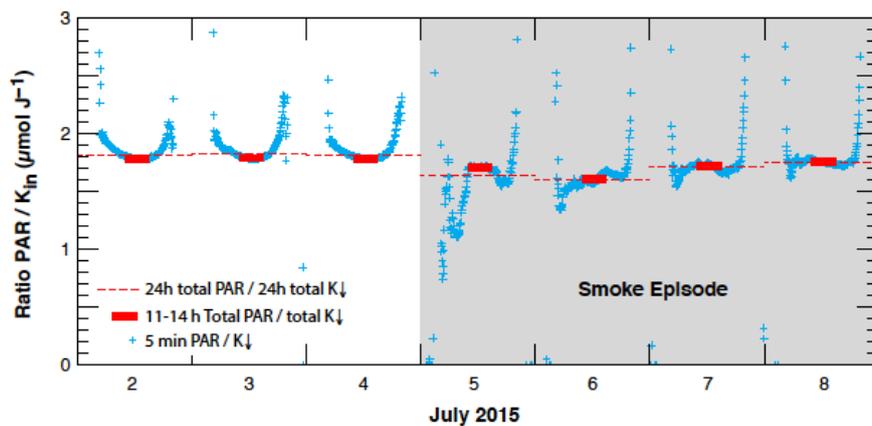


Figure 7: PAR/K_{\downarrow} ratio at Burns Bog from 2-8 July 2015

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