Effects of Crystal Phases and Morphology of Aluminum Oxide on the Heterogeneous Catalytic Ozonation Process Using Pyruvic Acid as an Indicator

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Aluminum oxide as an efficient heterogeneous catalyst has received much attention in the field of ozonation. In this study, the impact of the intrinsic properties of Al_2O_3 , including its crystalline phase and morphology, on the ozone-catalytic oxidation process were evaluated. The nanoparticles, nanotubes, nanorods, nanoleaves, nanoflakes and spindle-like 3D nanoarchitectures were constructed *via* hydrothermal methods and the crystal phases such as γ -, δ -, θ - and α - Al_2O_3 were controlled by appropriate calcinations using boehmite as precursor. The catalytic activity of the samples was tested by using degradation of pyruvic acid in aqueous solution in the presence of ozone, with a semicontinuous laboratory reactor. The initial degradation rates of pyruvic acid were found to correlate to the number of surface OH groups on the Al_2O_3 . Based on these findings, along with additional information from the techniques transmission electron microscopy, X-ray diffraction and Brunauer-Emmett-Teller measurements, it is concluded that large surface areas and the α -phase of Al_2O_3 facilitate the degradation of pyruvic acid with ozone. The morphology effects of the catalysts were negligible in present experiments.

Key Words: Heterogeneous catalytic ozonation, Al_2O_3 , Crystalline phase, Morphology, Pyruvic acid.

INTRODUCTION

Ozonation is a favoured method of water treatment due to its strong oxidizing ability, high solubility in water and leaving no harmful byproducts or sludges in the treated effluent¹. In the past decades, heterogeneous catalytic ozonation, which combines catalysts with ozone for enhanced oxidation of refractory organic species, began to attract widespread attention due to having various advantages over traditional ozonation methods. These advantages include enhanced ozone utilization, increased efficiency of pollutant removal and improved organic matter mineralization^{2,3}.

Aluminum oxide, or alumina, is a strong enough catalyst to facilitate the decomposition of organic compounds with ozone⁴. For example, alumina has been reported to catalyze the ozonation of 2-chlorophenol⁵ and also other organic acids such as acetic acid, oxalic acid, pyruvic acid, salicylic acid and succinic acid⁶⁻⁹. Hence, the catalytic capacity of pure alumina for the ozonation of organic compounds could be advantageous to investigate further.

The crystal phase has been found to play an essential role in determining the catalytic performance of alumina. Alumina can exist as 8 different polymorphs (γ , δ , ρ , η , θ , κ , χ and α -phase), each one having specific physicochemical properties.

Amongst the polymorphs, γ -, δ -, θ - and α - Al_2O_3 are always used in heterogeneous catalysis. The performance of various aluminum (hydroxyl) oxides, such as γ - $\text{Al}(\text{OH})_3$, γ - Al_2O_3 and α - Al_2O_3 , in decomposing aqueous ozone has been previously explored. The results show that the three aluminum oxides each exhibit a different catalytic activity, for the catalytic ozonation of 2,4,6-trichloroanisole¹⁰. However, this study did not concern the effect of specific surface area on the reaction rate. The surface area has been demonstrated to play an important role in the catalytic ozonation process¹¹. Thus, a systematic investigation into the influence of the crystal phase of alumina on catalytic ozonation should be made on the basis of subtracting the influence of the surface area.

The catalytic activities of materials strongly depend on their morphology, and this has been demonstrated in many fields¹²⁻¹⁴. In general, nanomaterials can be categorized into zero-, one-, two- and three-dimensional (0D, 1D, 2D and 3D) nanostructures. Alumina nanoparticles (0D) are an important material when used in heterogeneous catalysis, owing to their high surface area and mesoporous properties¹⁵. Recently, the use of one-dimensional (1D) nanostructured materials, such as nanowires, nanorods, nanobelts and nanotubes, are of great interest, since their novel properties originate from their unique structural one-dimensionality and quantum size confinement

effects in the other two dimensions¹⁶. In contrast to early studies which have focused on zero and one dimensional nanostructures, special attention has been recently paid to the synthesis of two-dimensional nanostructures¹⁷. Due to high anisotropy factors, some unique properties were observed in the two-dimensional nanostructures, which suggest potential to be future high performance catalysts^{17,18}. Moreover, the three-dimensional (3D) superstructures, which are constructed using nanoscale building blocks such as nanoparticles, nanorods, nanowires and nanosheets, may have an even greater potential for catalytic applications than the 1D nanostructures. Their potential applications could include adsorption, ceramics, photonics, electrochemistry and biomedicine¹⁹. Therefore, it is predicted that the shape of Al_2O_3 will have a strong influence on the catalytic ozonation process. However, there is little knowledge about what effect morphology has on catalytic ozonation processes.

The main aim of this work is to establish the interrelation between the morphology of alumina and the initial degradation rate (IDR) of the catalytic ozonation of pyruvic acid, as well as the crystal structure and the IDR, based on the quantitative determination of surface OH groups. Pyruvic acid was chosen as a model target pollutant in our study because it is a typical recalcitrant intermediate in the ozonation of dissolved aromatic compounds.

EXPERIMENTAL

In addition to the synthesis of nanoleaf using AlCl_3 as starting material, $\text{Al}(\text{NO}_3)_3 \cdot 9\text{H}_2\text{O}$ was chosen as Al precursor for preparing the rest Al_2O_3 catalysts. $\gamma\text{-Al}_2\text{O}_3$ nanoparticles were synthesized using the thermal decomposition of boehmite. δ -, θ - and α -nanoparticles were prepared by appropriately calcining $\gamma\text{-Al}_2\text{O}_3$ nanoparticles^{16,20-23}. Nanotubes, nanorods, nanoleaves, nanoflakes and spindle-like nano-alumina were prepared using hydrothermal processes, followed by different methods of calcinations^{16,17,24-26}.

The nomenclature used to represent the catalysts is as follows: X-Y-Z, where X represents the abbreviation of the shape of the catalyst, including nanoparticle (NP), nanotube (NT), nanorod (NR), nanoleaf (NL), nanoflake (NF) and spindle-like nanoarchitecture (NS), Y denotes the annealing temperature ($^\circ\text{C}$) and Z is the crystalline phase.

General procedure: To test the different catalytic activity of the prepared Al_2O_3 catalysts in the degradation of pyruvic acid, ozonation experiments were performed in a semicontinuous system (a continuous ozone gas supply with a fixed volume of water sample) by using the apparatus shown in Fig. 1. Ozone gas was generated from pure oxygen using a CHYF-3A ozone generator (Hangzhou Rongxin Electronic Equipment Co., Ltd., China). The flow rate of the mixture of O_2 and O_3 was regulated by a mass flow controller (D07-12A/ZM, Beijing Qixing Electronic Equipment Co., Ltd., China) at 300 mL/min. The mixture gas flow was produced as fine bubbles through a porous plastic diffuser located at the bottom of a flat-bottomed glass reactor, which had a total volume of 400 mL (height of 320 mm and an inner diameter of 40 mm). Excess ozone leaving the reactor was trapped by gas absorption bottles containing a 2% (w/v) KI solution. Before each

Fig. 1. Apparatus for ozonation: 1. Oxygen cylinder. 2. Ozone generator. 3. Mass flow controller. 4. Porous pellet. 5. Ozonation reactor. 6. Sampling system. 7. Magnetic stirrer. 8. KI solution. 9. Water bath vessel. 10. Thermometer

run, 300 mL of simulated wastewater containing 0.9 mmol pyruvic acid and 1.2 g of catalysts were introduced into the reactor. The reaction solution was stirred continuously using a magnetic stirring bar for good mixing. A sample of the suspension was withdrawn at preset time intervals to determine the residual chemical oxygen demand (COD). The solution temperature was controlled at $20 \pm 1^\circ\text{C}$ by a thermostatic bath (THD-2006, Ningbo Tianheng Instrument Factory, China) and the ozone concentration of the fed-gas was determined to be *ca.* 55 mg/L using a known iodimetric method²⁷. Measurement of the COD was performed according to standard methods. The same procedures were used for the control experiments of single ozonation without a catalyst.

It should be noted that pyruvic acid can be adsorbed by Al_2O_3 powders^{9,28}. To eliminate the effect of adsorption on the evaluation of catalytic activities of Al_2O_3 upon pyruvic acid degradation, COD values were measured without the removal of Al_2O_3 particles. This method is acceptable since no obvious differences were observed in the COD values with and without using Al_2O_3 powders in an additional experiment.

Detection method: X-ray diffraction (XRD), transmission electron microscopy (TEM) and Brunauer-Emmett-Teller (BET) were adopted to characterize the samples with respect to their phase composition, morphology and phase structure. The concentration of surface OH groups (expressed in mmol per unit gram) was measured by a saturated deprotonation method as described²⁹. The method is demonstrated to be accurate and the amount of the dissolved aluminum was found to be negligible.

RESULTS AND DISCUSSION

Catalyst characterization: For correlating the physico-chemical properties of the catalysts with the heterogeneous catalytic degradation of pyruvic acid in the presence of ozone, the Al_2O_3 powders were characterized by the means of XRD, TEM and BET techniques.

Assisted by using the XRD technique, we characterized the crystal phases of the catalysts to determine whether a single-phase aluminum oxide had been prepared. As shown in Fig. 2, all the reflections can be indexed to the alumina without

any indication of crystalline byproducts and different calcination routes lead to quite different crystalline phases. The diffraction peaks of the sample nanoparticle-500- γ can be assigned to a pure crystalline phase, with the lattice constants being in good agreement with $\gamma\text{-Al}_2\text{O}_3$ (indexed to JCPDS 10-0425). Using various temperature programs, the nanoparticle powders calcined at 900 °C could correspond to $\delta\text{-Al}_2\text{O}_3$ (JCPDS 04-0877) and $\theta\text{-Al}_2\text{O}_3$ (JCPDS 11-0517), respectively. Moreover, all detectable peaks of sample nanoparticle-1250- α belong to $\alpha\text{-Al}_2\text{O}_3$ (JCPDS 10-0173).

Fig. 2. XRD patterns of Al_2O_3 catalysts with various (a) morphology and (b) crystal phases

The grain size of the nanoparticle materials was roughly estimated using Scherrer's formula. Applied to the peaks of (440), (042), (240) and (113) of NP-500- γ , NP-900- δ , NP-900- θ and NP-1250- α , grain sizes of 4.4, 5.9, 6.0 and 177.2 nm were obtained, respectively. This implies the crystalline size increases with an increased annealing temperature.

In addition, NP-500- γ , NT-500- γ , NR-600- γ , NL-600- γ , NF-600- γ and NS-550- γ all had similar 2θ values of the diffraction peaks, exhibiting clear diffraction peaks at 19.3° , 31.9° , 37.2° , 39.4° , 45.8° , 60.8° and 66.9° . These values could be assigned to the well crystallized γ -phase of (111), (220), (311), (222), (400), (511) and (440) crystal planes, respectively, when compared to standard diffraction peaks (JCPDS 10-0425). No other peak was observed belonging to the impurities. This suggests that no additional phase was contained in the samples except the γ forms.

TEM analysis has been conducted to clearly depict the characteristics of the samples. Fig. 3 a and g-i show that with an increase in the calcination temperature from 500 to 1250 °C, the particle diameter of the aluminum oxides increases from *ca.* 5 to *ca.* 130 nm. This is roughly in accordance with the Scherrer analysis of the XRD data. The enlargement of the catalyst nanocrystals should be the consequence of the agglomeration of the phase present. Furthermore, it can be clearly seen that the morphologies are different from those of the samples prepared using different routes. As reflected in Fig. 3, the samples of NP-500- γ , NT-500- γ , NR-600- γ , NL-600- γ , NF-600- γ and NS-550- γ all display the clear-cut images of nanoparticle, nanotube, nanorod, nanoleaf, nanoflake and spindle-like morphology, respectively. The nanotubes have an outer diameter of *ca.* 5.5 nm and an inner diameter of *ca.* 2 nm. The straight nanorods have diameters ranging from 12 to 15 nm, with lengths of about 100-300 nm. The shapes of the plate-like nanostructures are quadrangles with sizes of 20-60 nm in width. The uniform leaf-like morphology shows a lateral size of approximately $4.0\ \mu\text{m} \times 8.5\ \mu\text{m}$ and the spindle-like particles are *ca.* 1.5 μm in length and are composed of many oriented nanoplatelets.

The BET specific surface area (S_{BET}) of the prepared samples is listed in Table-1. It can be observed that the S_{BET} generally decreases with an increase in the annealing temperature. S_{BET} values are 231, 121, 137 and $11.8\ \text{m}^2/\text{g}$ and correspond to the crystalline phases of γ , δ , θ and α , respectively. In addition, samples annealed in the temperature range of 500-600 °C show largely different specific surface areas as a function

TABLE-1
REFINED PHYSICOCHEMICAL PARAMETERS AND THE INITIAL DEGRADATION
RATE OF PYRUVIC ACID (PA) FOR THE PREPARED CATALYSTS

Sample	S_{BET} ($\text{m}^2\ \text{g}^{-1}$)	Crystallite phase	Concentration of surface-OH ($\text{mmol}\ \text{g}^{-1}$)	Density of surface- OH ($\mu\text{mol}\ \text{m}^{-2}$)	Initial degradation rate (IDR) ($\text{mg}\ \text{L}^{-1}\ \text{min}^{-1}$)	IDR/ S ($\text{mg}\ \text{L}^{-1}\ \text{min}^{-1}\ \text{m}^{-2}$)
NP-500- γ	231	γ	0.247 ± 0.014	1.07 ± 0.06	8.31 ± 1.04	0.036 ± 0.005
NT-500- γ	214	γ	0.245 ± 0.005	1.15 ± 0.02	6.18 ± 0.87	0.029 ± 0.004
NR-600- γ	223	γ	0.211 ± 0.010	0.95 ± 0.05	7.07 ± 0.80	0.032 ± 0.004
NL-600- γ	58	γ	0.064 ± 0.005	1.10 ± 0.09	2.32 ± 0.58	0.040 ± 0.010
NF-600- γ	89	γ	0.096 ± 0.001	1.08 ± 0.01	3.49 ± 0.33	0.039 ± 0.004
NS-550- γ	205	γ	0.212 ± 0.011	1.04 ± 0.05	5.34 ± 0.44	0.026 ± 0.002
NP-900- δ	121	δ	0.160 ± 0.004	1.32 ± 0.04	6.11 ± 0.24	0.051 ± 0.002
NP-900- θ	137	θ	0.188 ± 0.010	1.37 ± 0.08	6.74 ± 1.77	0.049 ± 0.013
NP-1250- α	11.8	α	0.033 ± 0.005	2.36 ± 0.28	1.15 ± 0.24	0.106 ± 0.020

Fig. 3. TEM images of catalysts powders: (a) γ - Al_2O_3 nanoparticles, (b) γ - Al_2O_3 nanotubes, (c) γ - Al_2O_3 nanorods, (d) γ - Al_2O_3 nanoleaves, (e) γ - Al_2O_3 nanoflakes, (f) spindle-like γ - Al_2O_3 3D nanoarchitectures, (g) δ - Al_2O_3 nanoparticles, (h) θ - Al_2O_3 nanoparticles, (i) α - Al_2O_3 nanoparticles

of morphology. The largest S_{BET} value of $231 \text{ m}^2/\text{g}$ obtained for the sample NP-500- γ , while the NL-600- γ has the smallest BET surface area of $58 \text{ m}^2/\text{g}$.

Activity of Al_2O_3 in the catalytic ozonation: A COD test is an effective tool for indirectly measuring the amount of organic compounds in water. Hence, COD reduction was used as the primary target for evaluating the efficiency of the catalytic ozonation of pyruvic acid. As shown in Fig. 3, ozone alone does not cause a significant degradation of pyruvic acid after 2 h of treatment because there was only about a 7.5 % decrease in COD with ozone alone. In comparison, catalytic ozonation with Al_2O_3 exhibited excellent performance. The difference between COD removal in the catalyzed and uncatalyzed ozonation processes indicated that the ozone itself

is not a strong enough catalyst to degrade pyruvic acid and alumina can enhance the ozone oxidation process.

It is well accepted that the degradation rate would fluctuate during each experimental run. This can be explained due to the complicating effects of reaction products undergoing further reactions. Accordingly, we used the initial degradation rate to evaluate the catalytic activity of Al_2O_3 . Initial degradation rate is the instantaneous rate at the beginning of a reaction. The solution pH was almost the same in the case for all experiments. In addition, the efficient degradation of pyruvic acid with ozone is believed to originate mostly from the surface OH groups on alumina, which will be discussed further. Therefore, it was not found necessary to control the pH of the solution in our experiments.

As depicted in Fig. 4 and Table-1, 9 types of alumina exhibited distinct remarkable catalytic activity in catalytic ozonation for the degradation of pyruvic acid. The catalytic activities assessed by the initial degradation rate decreased in the following order: NP-500- γ > NR-600- γ > NP-900- θ > NT-500- γ > NP-900- δ > NS-550- γ > NF-600- γ > NL-600- γ > NP-1250- α . At first inspection, the calcined temperature is shown to have a negative effect on the catalytic activity, as the pyruvic acid degradation decreases with the increasing temperature of calcination. Furthermore, activities of alumina seem to be controlled by the morphology of the catalysts. Nanoparticles calcined in air at 500 °C (*i.e.*, NP-500- γ) reveal the highest pyruvic acid conversion whereas nanoparticles calcined in air at 1250 °C are the least efficient catalyst. The initial degradation rate of the former reached 8.31 mg/(L min), which exceeded that of the later by a factor of about 7.23-fold.

Fig. 4. COD removal during the degradation of pyruvic acid by single ozonation and by ozonation over Al_2O_3 with various morphology (a), as well as by ozonation over Al_2O_3 with various crystal phases. Experimental conditions: initial concentration of pyruvic acid, 3 mmol/L; catalyst dose, 4.0 g/L; O_3/O_2 flow rate, 300 mL/min

Álvarez *et al.*⁹ reported that the removal of pyruvic acid by ozonation in the presence of Al_2O_3 occurred on the catalyst surface, where the carboxylate ions of the acids substitute the surface hydroxyl groups of alumina^{28,30}. Thus, for investigating the effect of the crystalline phases and morphology of the catalysts on the catalytic activities of alumina, surface OH groups were determined accurately and the results are given in Table-1. It is clear from Fig. 5 that there is a strong linear correlation between initial degradation rate and the relative

concentration of surface OH in the degradation of pyruvic acid with ozone. On this basis, the degradation trend can be interpreted in detail from the effects of crystal phases and morphology.

Fig. 5. Relationship between the initial degradation rate (IDR) of pyruvic acid (PA) and the concentration of surface OH groups on catalysts

Effect of morphology on the catalytic ozonation of pyruvic acid: The morphology of catalysts may affect the catalytic activities. In relation to the effect of morphology on the catalyzed ozonation of pyruvic acid, various shapes of γ - Al_2O_3 , referred to as NP-500- γ , NT-500- γ , NR-600- γ , NL-600- γ , NF-600- γ and NS-550- γ , respectively, have been recorded in Fig. 3a and Table-1. The examined catalyzed ozonation data exhibited varied degrees of degradation, with the initial degradation rate ranging from *ca.* 2.32 to 8.31 mg/(L min). Also, Al_2O_3 with a higher specific surface area displayed a faster initial degradation rate, with values of 8.31, 7.07, 6.18, 5.34, 3.49 and 2.32 mg/(L min) when the BET area is 231, 223, 214, 205, 89 and 58 m^2/g , respectively. However, no obvious differences of the IDR/S could be found among different morphology of γ - Al_2O_3 including nanotubes, nanorods, nanoleaves, nanoflakes and spindle-like nanoarchitecture.

Similarly, Table-1 shows a trend that the concentration of surface OH strongly depends on the surface area for samples with the γ phase. The data suggest that a higher concentration of surface OH is featured on the higher S_{BET} of catalyst and appears propitious to the degradation of pyruvic acid. This implies that the surface area is a major factor in achieving high catalytic reactivity of ozone. Nevertheless, using the comparison method whereby the effect of surface area is subtracted, we found from Fig. 6a and 6b that IDR/S and the number of surface OH per unit area (called density of surface OH) is not marked enough for Al_2O_3 nanostructures with various morphology. In other words, their surface hydroxyl content makes no obvious difference in the unit area. Thus it is suggested that morphology of alumina has very little influence on the alumina-catalyzed ozonation of pyruvic acid.

Impact of crystal phase of Al_2O_3 on heterogeneous catalytic ozonation of pyruvic acid: As reported, all phases of Al_2O_3 have defects. γ - and δ - Al_2O_3 are aluminum deficient, whereas θ - and α - Al_2O_3 are deficient in oxygen atoms³¹⁻³⁴.

Fig. 6. Comparison of IDR/S (a) and the density of surface OH groups on catalysts (b) for Al_2O_3 with various crystalline phases and morphology

Vacancy defects in the crystalline structures of solid catalysts, both cationic and anionic in the crystalline structures, can act as active reaction centers *via* the generation of surface OH groups, thus determining the rate of the heterogeneous ozonation process³⁵. To study the effects of phase structures of the alumina nanoparticles on the catalytic activities, rates of ozonation of pyruvic acid with respect to γ -, δ -, θ - Al_2O_3 and α - Al_2O_3 were examined. The four samples used in the catalytic activity test were morphologically very similar but presented different surface areas. As shown in Fig. 4b, the α - Al_2O_3 catalyzed process gave only a 15.4 % decrease in COD after 2 h of reaction. In comparison, the COD removal is 56.0, 40.3 and 46.7 % for the γ -, δ - and θ - Al_2O_3 nanoparticles, respectively. The initial degradation rate of pyruvic acid degradation followed a very similar trend. In particular, the initial degradation rate of pyruvic acid degradation with NP-1250- α was approximately one-seventh smaller than that of the NP-500- γ . Considering that the BET surface areas can greatly influence the catalytic activity as mentioned above, the BET surface area should be subtracted from initial degradation rate to establish the effect of the nature of the crystalline phase on the catalytic ozonation of pyruvic acid. It is obvious from Fig. 6 and Table-1 that the IDR/S increased in the following order: γ - $\text{Al}_2\text{O}_3 < \delta$ - $\text{Al}_2\text{O}_3 \approx \theta$ - $\text{Al}_2\text{O}_3 < \alpha$ - Al_2O_3 and the same trend can be found for the yield in the density of surface OH.

As aforementioned, differences in catalytic activity amongst the various Al_2O_3 catalysts depend on the nature of vacancies. In our experiments, using calcination techniques,

we converted boehmite to transition γ -, δ -, θ - Al_2O_3 and to finally to α - Al_2O_3 . It is widely accepted that the high-temperature transition phases are less active than the low temperature ones due to the lower surface area of the former. This effect is brought about by the higher order and larger particle size. Simultaneously, there exists a larger fraction of tetrahedrally coordinated Al^{3+} ions in the low-temperature transition alumina than in high-temperature transition alumina³¹. However, if we normalized the surface area to account for the relationship between the crystalline phase and surface OH groups in this work, values of the density of surface OH groups was found to be decreasing in the order of α - $\text{Al}_2\text{O}_3 > \theta$ - $\text{Al}_2\text{O}_3 \approx \delta$ - $\text{Al}_2\text{O}_3 > \gamma$ - Al_2O_3 . The α - Al_2O_3 appears to be significantly higher than other phases. The following reason can be formulated to explain this sequence. The generation of surface OH groups in aqueous solution is different from that under ambient conditions. It is well known that hydrated alumina surfaces undergo phase transformations during the hydration process^{33,36}. For example, the surface of γ - Al_2O_3 can hydrate to bayerite with prolonged hydration and a mixture of α - $\text{Al}(\text{OH})_3$ and γ - $\text{Al}(\text{OH})_3$ is produced when the surface of α - Al_2O_3 is covered with water³⁷⁻³⁹. In this case, the proportion of singly and doubly coordinated aluminum-oxygen bonding contributes to the amount of surface OH groups. Therefore, we can deduce that Al_2O_3 has various numbers of hydroxyl groups in aqueous solution according to the crystalline phase. α - Al_2O_3 possess the highest density of surface OH groups, followed by θ -, δ - and finally γ -. From the above information it is concluded that the crystal phase plays a vital role in the catalytic ozonation process and the α - Al_2O_3 phase is more beneficial to the system.

Conclusion

To summarize, the catalytic activity of Al_2O_3 for the oxidation of pyruvic acid with ozone is dominated by surface hydroxyl groups which are dependent mainly upon their specific surface areas and crystal phase, whereas the effect of morphology on the catalytic performance seems to be negligible. A high-activity catalyst requires simultaneously maximizing the surface area and the content of the α -form of Al_2O_3 .

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